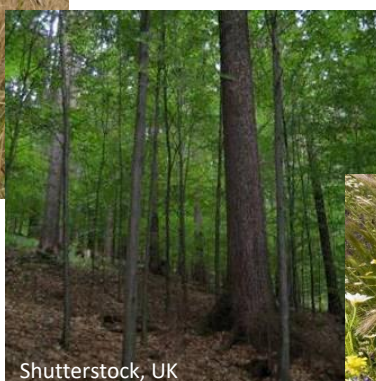


SCIENTIFIC BACKGROUND DOCUMENT A

Version October 2018

SUPPLEMENT OF CHAPTER III (MAPPING CRITICAL LEVELS FOR VEGETATION) OF THE MODELLING AND MAPPING MANUAL OF THE LRTAP CONVENTION



Chapter 3 of the Modelling and Mapping Manual of the LRTAP Convention describes the most up-to-date methodology and establishment of critical levels for adverse impacts of air pollutant (ozone, sulphur dioxide, nitrogen oxides and ammonia) on vegetation. The current version of Chapter 3 includes updates to critical levels for ozone agreed at the 30th ICP Vegetation Task Force Meeting, 14-17 February, 2017, Poznan, Poland.

For ozone, further supporting information is provided in this Scientific Background Document A (SBD-A) regarding the methodologies and critical levels described in Chapter 3. In addition, a Scientific Background Document B (SBD-B) is available, containing DO₃SE model parameterisations for additional species and information on developing areas of ozone research and the application of methodologies to further develop ozone critical levels. Chapter 3 and both scientific background documents are available on the ICP Vegetation website at <http://icpvegetation.ceh.ac.uk>.

* International Cooperative Programme on Effects of Air Pollution on Natural Vegetation and Crops

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Please refer to this document as: ICP Vegetation (2017). Scientific Background Document A of Chapter 3 of 'Manual on methodologies and criteria for modelling and mapping critical loads and levels of air pollution effects, risks and trends', UNECE Convention on Long-range Transboundary Air Pollution. Accessed on [date of consultation] at <http://icpvegetation.ceh.ac.uk>.

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1 Introduction

This Scientific Background Document A (SBD-A) contains supplementary information for Chapter 3 ('Mapping critical levels for vegetation') of the Modelling and Mapping Manual of the Convention on Long-range Transboundary Air Pollution (LRTAP), specifically for ozone (O₃) critical levels for vegetation. Chapter 3 of the manual was prepared under the leadership of the ICP Vegetation and was fully revised to include updates to critical levels for O₃ agreed at the 30th ICP Vegetation Task Force Meeting, 14-17 February, 2017, Poznan, Poland. The revised version of Chapter 3 was published in April 2017 on the ICP Vegetation website (<http://icpvegetation.ceh.ac.uk>). Where relevant, reference is made within brackets to the appropriate sections of Chapter 3.

SBD-A contains further information on (with reference to relevant Section in the manual):

- Method for setting a reference value (Ref10 POD_y) for determining flux-based critical levels (Section III.3.1.3);
- Modelling the O₃ concentration at the top of the canopy (Section III.3.4.2);
- Scientific bases of the parameterisation of the O₃ flux models used to establish critical levels for crops (Section III.3.5.2), forest trees (Section III.3.5.3) and (semi-) natural vegetation (Section III.3.5.4);
- Scientific bases of O₃ concentration-based (AOT40) dose-response relationships and critical levels for crops, forests trees and (semi-)natural vegetation (Section III.3.7);
- O₃-sensitivity of plant species.

In the previous version of Chapter 3 of the manual, much of the information related to bullet points two and three was available in the Annexes of Chapter 3.

2 Method for setting a reference value (Ref10 POD_Y) for determining flux-based critical levels (Section III.3.1.3 of the manual)

2.1 Introduction of the methodology

At critical levels expert workshop in Deganwy, UK (7 – 9 June, 2016) and at the UNECE Ozone Critical Levels Workshop in Madrid, Spain (7 – 8 November, 2016) concern was raised that in some circumstances when the stomatal O_3 flux approach is being used and the Phytotoxic Ozone Dose above a flux-threshold of Y (POD_Y) is being calculated, reference against a POD_Y of 0 could theoretically lead to an O_3 critical level that is not achievable as the O_3 concentrations needed to achieve this could be lower than 'pre-industrial' O_3 . This is especially the case when Y is low. Hence, the calculation of a reference POD_Y (Ref POD_Y) was explored in setting critical levels for vegetation. **At the workshop in Madrid it was decided that the reference POD_Y should be calculated at constant 10 ppb O_3 (Ref10 POD_Y), representing clean air, by:**

- Calculating the Ref POD_Y by using a constant O_3 concentration of 10 ppb and the climatic conditions in the experiment;
- Where data is combined from different climates, take the mean of this POD_Y as the reference POD_Y from which you calculate the critical level.

The application of this approach is shown in **Figure 2.1**.

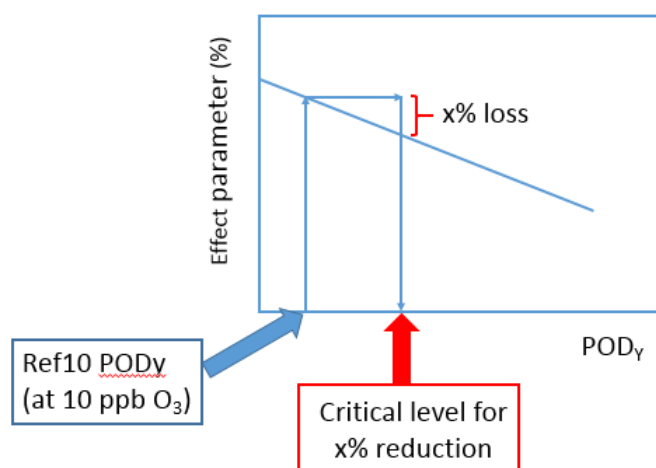


Figure 2.1 Method for using Ref10 POD_Y (i.e. POD_Y at 10 ppb constant O_3) as reference point for O_3 critical level derivation.

In this paper we provide further details on justification for the introduction of the above methodology, using a constant O_3 concentration of 10 ppb, representing clear air and potentially 'pre-industrial' O_3 concentrations.

2.2 Evaluation of 'pre-industrial' O_3

The only quantitative measurements of O_3 during the 19th century were made in Park Montsouris on the outskirts of Paris. Volz and Kley (1988) conducted a reanalysis of the Montsouris data and found that the early technique produced results similar to a modern ultraviolet absorption instrument. They concluded that average O_3 at Montsouris during 1876–1910 was 11 ppbv with an uncertainty of ± 2 ppbv. However, there is no way to know if these values were representative of other surface locations in the Northern Hemisphere. Measurements conducted in the latter half of the 19th century at various locations across the world using the Schönbein paper methodology are deemed to be very uncertain with

respect to absolute O₃ values (Cooper et al., 2014). Despite the uncertainty in absolute values, these measurements showed that 1) 19th century seasonal O₃ most often peaked in spring, followed by winter; seasonal O₃ minima most frequently occurred in summer and autumn, and 2) studies that compared late 19th century estimated O₃ to late 20th century ultraviolet absorption O₃ measurements generally concluded that average O₃ increased by about a factor of two (Cooper et al., 2014) or more (Anfossi et al., 1991; Volz and Kley, 1988) during the 20th century.

Despite the uncertainty of the Schönbein method, there is one set of late 19th century Schönbein measurements that is difficult to dismiss. Measurements made between 1874 and 1909 at Pic du Midi, France, 3000 m above sea level, used the same type of paper and techniques as those employed at Montsouris (Marenco et al., 1994). Accounting for differences in pressure and humidity between Pic du Midi and Montsouris, Marenco et al. (1994) used the Montsouris regression to estimate that Pic du Midi O₃ concentrations were approximately 10 ppbv during 1874–1895, with a springtime peak and wintertime minimum. From 1895 until the end of the record in 1909 O₃ increased steadily to 14 ppbv, while O₃ at Montsouris at this time decreased. Marenco et al. (1994) point out that the increase in O₃ at Pic du Midi coincided with global increases in methane concentrations, while the O₃ decrease at Montsouris may have been the result of increased emission of NO in Paris that reduced O₃ concentrations by titration at nearby Montsouris. While the Pic du Midi O₃ mixing ratios appear to be more reliable than any other record outside of Montsouris, the results raise the question why O₃ at 3 km above sea level at Pic du Midi, a site representing the free troposphere most of the time, was no greater than O₃ at the low elevation site of Montsouris. This lack of a vertical O₃ gradient in the lower troposphere is in direct contrast with O₃ sonde observations from around the world that showed a consistent increase in O₃ with altitude during the 1980s and 1990s (Logan, 1999). The latest generation of atmospheric chemistry models taking part in the Atmospheric Chemistry and Climate Model Intercomparison Project (ACCMIP) show a vertical O₃ gradient in the northern mid-latitude lower troposphere during the 1850s that is weaker than the period 1996–2005, but the models overestimate 1850s lower tropospheric O₃ by a factor of two compared to Pic du Midi and Montsouris (Stevenson et al., 2013). ACCMIP estimated surface O₃ in the northern hemisphere mid-latitudes in 1850 to be more than 40% lower than in 2000, with absolute decreases more than 25 ppbv for the Mediterranean, much of Asia, and the western USA due to less precursor emissions (Young et al., 2013).

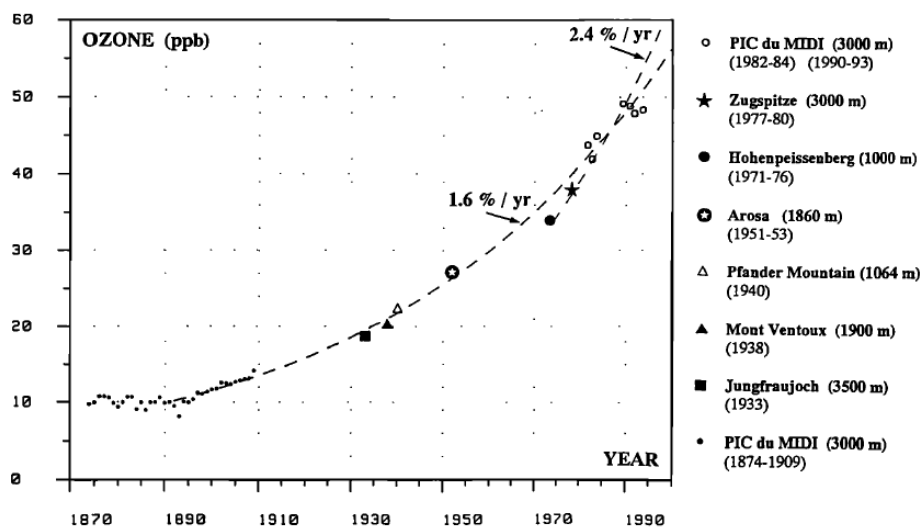


Figure 2.2 O₃ evolution in the free atmosphere over Western Europe, from measurement at the Pic du Midi and various European stations at high altitudes. Source: Marenco et al. (1994).

Marenco et al. (1994) showed an exponential increase in O₃ across Western Europe using measurements from several locations, beginning with the Pic du Midi Schönbein measurements in the 1870s followed by short-term quantitative measurements **at several high altitude sites** in Switzerland, Germany, and France from the 1930s through the early 1990s (**Figure 2.2**). They concluded that O₃ increased by a factor of 5 between the late 1800s and the early 1990s and by a factor of 2 between the 1950s and early 1990s. This increase corresponds to the global increase in fossil fuel combustion summarized by the IPCC Fifth Assessment Report (IPCC, 2013).

2.3 Estimated range of O₃ concentrations early 1900s

Table 2.1 provides an estimate of surface O₃ concentrations in the Northern Hemisphere based on the concentrations in 2000 for sites at different elevations (Cooper et al., 2014) and assuming a rise in surface O₃ concentration of a factor 2 (Cooper et al., 2014), 3 (Marenco et al., 1994) or 5 (Marenco et al., 1994). Equally, the rise in surface O₃ concentration by a factor 2 could also reflect the ground-level O₃ concentration in the 1950s as O₃ concentrations might have doubled since then (Cooper et al., 2014; Marenco et al., 1994). The average surface O₃ concentration at low and high elevation sites agrees well with measurements at Montsouris and Pic du Midi (see introduction) when assuming a rise in surface O₃ concentration by a factor 3 between 1900 and 2000. The Royal Society (2008) came to the conclusion that a background O₃ concentration of 10 – 15 ppbv in 1900 had doubled to 20 – 30 ppbv by 1980 and has since then risen by another 5 ppbv till 2007. The world’s longest continuous O₃ record is from the Arkona-Zingst site on the northern German coast. In the late 1950s and early 1960s annual mean O₃ concentrations were between 15 – 20 ppbv, which had doubled by the end of the 20th century (Cooper et al., 2014).

Table 2.1 *Estimate range of ground-level O₃ concentration in 1900 at sites in the Northern Hemisphere, based on the O₃ concentration at the sites in 2000 (Cooper et al., 2014), and assuming a linear rise in ground-level O₃ concentration by a factor 2, 3 and 5 between 1900 and 2000.*

Sites	Elevation (m)	Rise since 1900	Factor 2	Factor 3	Factor 5
		Year 2000 Ozone (ppbv)	1900 Ozone (ppbv)	1900 Ozone (ppbv)	1900 Ozone (ppbv)
Barrow, Alaska	0	27	13	9	5
Arkona-Zingst, Germany	0	28	14	9	6
Storhofdi, Iceland	100	38	19	13	8
Japanese MBL	100	44	22	15	9
Mace Head, Ireland	200	39	19	13	8
US Pacific MBL	200	32	16	11	6
Hohenpeissenberg, Germany	1000	41	21	14	8
Arosa, Switzerland	1800	45	22	15	9
Lassen NP, California, USA	1800	40	20	13	8
Mt Happa, Japan	1900	51	25	17	10
Zugspitze, Germany	3000	53	27	18	11
Summit, Greenland	3200	45	23	15	9
Mauna Loa, Hawaii, USA	3400	40	20	13	8
Jungfrauojoch, Switzerland	3600	54	27	18	11
Average	All	41	21	14	8
	≤200 m	35	17	12	7
	≥1000 m	46	23	15	9

2.4 Examples of Ref POD_Y calculations for different receptors and using constant O_3 concentrations in the range of 10 – 25 ppb

The DO_{3SE} (Deposition of O_3 for Stomatal Exchange) model (<https://www.sei-international.org/do3se>) was used to calculate stomatal O_3 fluxes at a constant O_3 concentration of 10, 15, 20 or 25 ppb, applying either the parameterisations as defined in the Modelling and Mapping Manual of the LRTAP Convention (LRTAP Convention, 2017) or local parameterisations where appropriate. Stomatal O_3 fluxes, i.e. Phytotoxic Ozone Doses (PODs), were calculated for different vegetation types, based on experiments conducted in various years under varying climatic conditions. For wheat and potato, data used in the establishment of the flux-effect relationships as defined in the Modelling and Mapping Manual were used. For trees, data described by B ker et al. (2015) were used. In addition, POD_Y was calculated for different plant species for 2010 using climate data from a gradient of seven sites across Europe, i.e.  stad (SE), Cairngorm, Auchencorth, Harwell (all UK), Giessen (DE), Arconate (IT), Tres Cantos (ES). The resulting O_3 stomatal fluxes are referred to as reference POD_Y (Ref POD_Y). The average reference POD_Y for the different plant species or plant functional types are shown in **Figure 2.3 and 2.4**.

As to be expected, Ref POD_Y is lowest for crops (Ref $POD_Y = 0$) as the flux threshold Y is $6 \text{ nmol m}^{-2} \text{ s}^{-1}$ and the accumulation period is shorter. Ref POD_Y is higher when using a generic crop flux model for application in integrated assessment modelling (Crop IAM, based on wheat parameterisation) as Y for this model is $3 \text{ nmol m}^{-2} \text{ s}^{-1}$ and the accumulation period is longer (LRTAP Convention, 2017). Compared with crops, Ref POD_Y is higher for (semi-)natural vegetation and tree species due to the low Y value of $1 \text{ nmol m}^{-2} \text{ s}^{-1}$ and the longer accumulation period (LRTAP Convention, 2017). Ref POD_Y increases linearly with reference O_3 concentration ($R^2 > 0.97$). Ref POD_Y is generally lower for needle leaf trees than deciduous broadleaf trees; remarkable are the high Ref POD_Y values for poplar, more than a factor two higher than for other deciduous broadleaf tree species, primarily due to its higher maximum stomatal conductance.

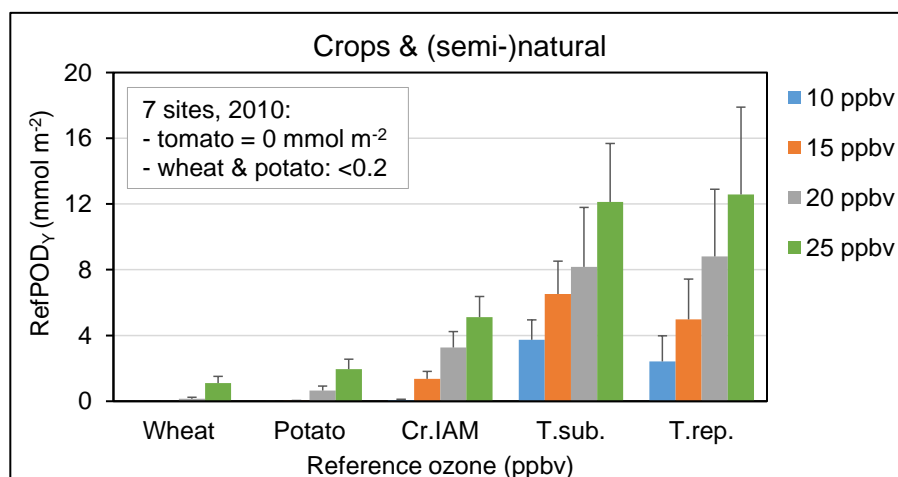


Figure 2.3. Reference Phytotoxic Ozone Dose (POD_Y) calculated at a constant O_3 concentration ranging from 10 to 25 ppbv for crops and (semi-)natural vegetation. $Y = 6, 6, 3$ and 1 for wheat, potato, generic crop (Cr.IAM) and Trifolium species ($T.sub.$ = Trifolium subterraneum; $T.rep.$ = Trifolium repens) respectively. $T.rep.$ values are an average for seven sites in 2010. Bars show mean + one SD.

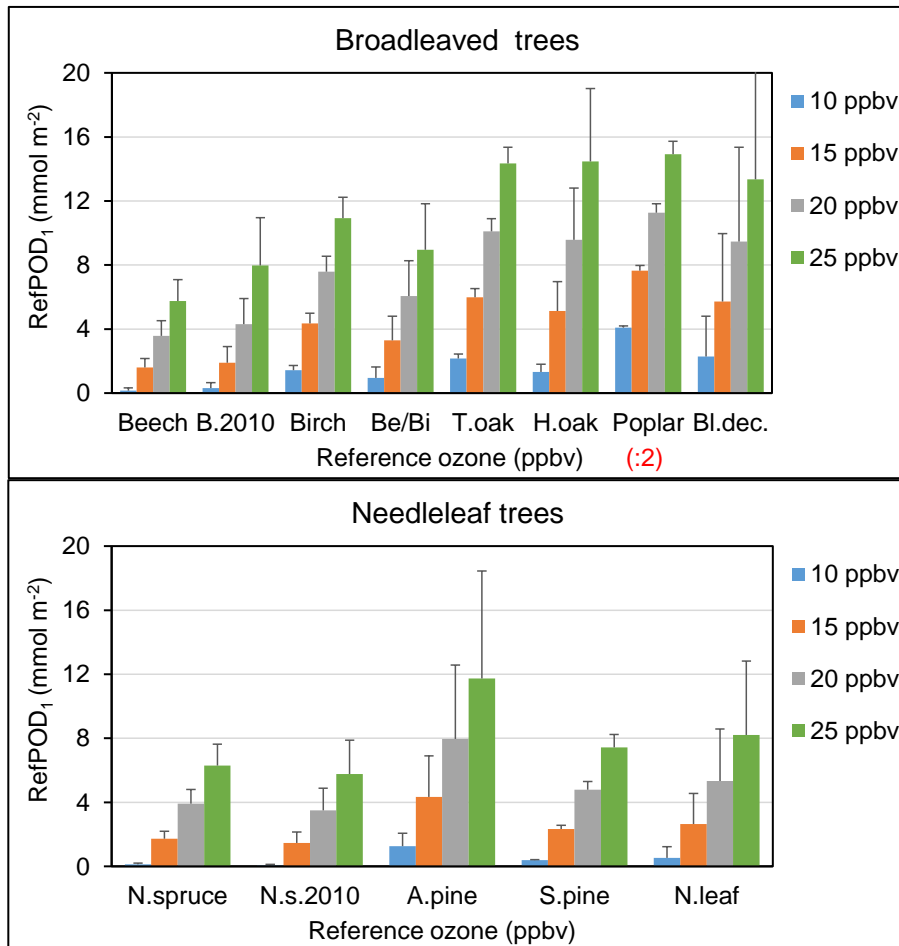


Figure 2.4 Reference Phytotoxic Ozone Dose (POD_{γ}) calculated at a constant O_3 concentration ranging from 10 to 25 ppbv for trees; $Y = 1$ for all tree species. Bars show mean + one SD. Top: Broadleaf deciduous tree species (B.2010 = Beech for seven sites in 2010; Be/Bi = Beech/Birch; T.oak = temperate oak; H.oak = Holm oak; Bl.dec. = all Broadleaved deciduous trees; **values for poplar were divided by two**). Bottom: Needleleaf tree species; N.spruce = Norway spruce; N.s.2010 = Norway spruce for seven sites in 2010; A.pine = Aleppo pine; S.pine = Scots pine; N.leaf = all needleleaf trees.

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3 Modelling the O₃ concentration at the top of the canopy (Section III.3.4.2 of the manual)

3.1 Introduction

The O₃ concentrations needed for the calculation of the POD and the critical levels are the concentrations which occur just at the upper limit of the laminar boundary layer of the receptors' leaves. Within exposure systems such as open-top chambers, where air flow is omnidirectional, the exposure concentration measured at the top of the canopy reflects the O₃ concentration at the upper boundary of the leaves. Under unenclosed field conditions, it was decided that the O₃ concentration at the top of the canopy provides a reasonable estimate of the O₃ concentration at the upper surface boundary of the laminar boundary layer near the flag leaf (in the case of wheat) and the sunlit upper canopy leaves (in the case of other receptors). Thus, the O₃ concentration at the top of the canopy should be available for determining each of the indices used.

Unfortunately O₃ concentration is generally not measured at the top of the canopy but well above it in the case of crops, or well below it in the case of forests (DO3SE, 2014). For crops and other low vegetation, canopy-top O₃ concentrations may be significantly lower than those at conventional measurement heights of 2-5m above the ground (**Figure 3.1a**), and hence use of measured data, directly or after spatial interpolation, may lead to significant overestimates of O₃ concentrations and hence of the degree of exceedance of critical levels. In contrast, for forests, measured data at 2-5m above the ground may underestimate O₃ concentration at the top of the canopy (**Figure 3.1b**). The difference in O₃ concentration between measurement height and canopy height is a function of several factors, including wind speed, atmospheric stability (note how in **Figure 3.1** the O₃ concentration profile changes with atmospheric stability) and other meteorological factors, canopy height, surface roughness and the total flux of O₃, F_{tot}.

Conversion of O₃ concentration at measurement height to that at canopy-top height (Z_{igt}) can be best achieved with an appropriate deposition model. It should be noted, however, that the flux-gradient relationships these models depend on are not strictly valid within the roughness sublayer (up to 2-3 times canopy height) or in a heterogeneous landscape, so even such detailed calculations can provide only approximate answers. The model chosen will depend upon the amount of meteorological data that is available.

3.2 Conversion of O₃ concentration at measurement height to canopy height

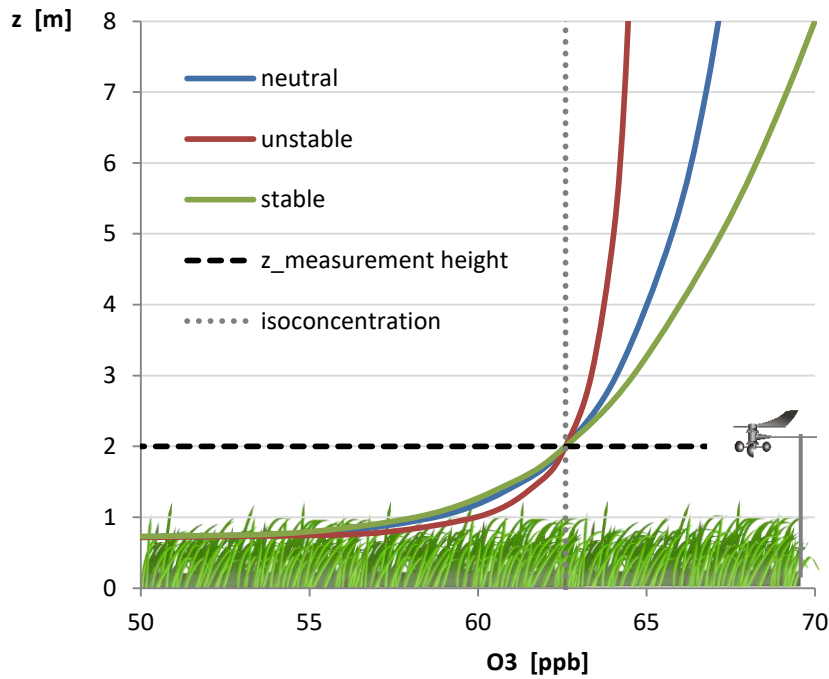
Three methods are included here which can be used to achieve the necessary conversion if (a) no meteorological data are available at all, if (b) all the important meteorological measurements are available, or if (c) some basic measurements are available.

Method a) Tabulated gradients

If no meteorological data are available at all, then a simple tabulation of vertical O₃ gradients can be used. The relationship between O₃ concentrations at a number of different heights has been estimated with the EMEP deposition module (Emberson et al., 2000b), using meteorology from about 30 sites across Europe. Data were produced for an arbitrary crop surface and for short grasslands and forest trees (see Table III.7 in Modelling and Mapping Manual). For the crop surface, the assumptions made here are that we have a 1 m high crop with g_{max} = 450 mmol O₃ m⁻² PLA s⁻¹. The total leaf surface area index (LAI) is set to 5 m² PLA m⁻², and the green LAI is set at 3 m² PLA m⁻², assumed to give a canopy-scale phenology factor (f_{phen}) of 0.6. The soil moisture factor (f_{sw}) is set to 1.0. Constant values of these parameters are used throughout the year in order to avoid problems with trying to estimate growth stage in different areas of Europe. The concentration gradients thus derived are most appropriate to a fully developed crop but will serve as a reasonable

approximation for the whole growing season. Other stomatal conductance modifiers are allowed to vary according to the wheat-specific functions.

a)



b)

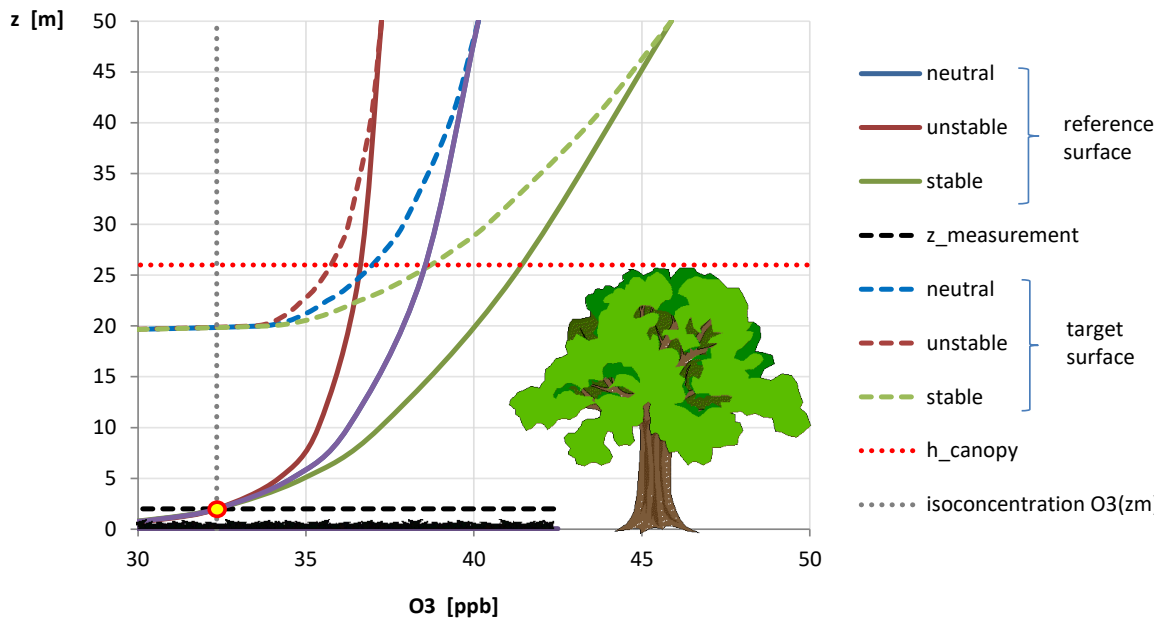


Figure 3.1

O₃ gradients with different atmospheric stability conditions (neutral, stable, unstable). a) Gradients above crops when O₃ concentration is measured above the canopy top; b) gradients above a forest (target surface) when O₃ concentration is measured above a reference surface with lower vegetation away from the forest edge at an height which is lower than the forest top-canopy height . Thick curves are the O₃ gradients above the reference surface generated from the measuring point (yellow/red dot); dashed curves are the O₃ gradients above the forest canopy.

For short grasslands, canopy height was set to 0.1 m, g_{\max} to 270 mmol O₃ m⁻² PLA s⁻¹ and f_{SWP} set to 1.0. All other factors are as given for grasslands in Emberson et al. (2000b). For the micrometeorology, the displacement height (d) and roughness length (z_0) are set to 0.7 and 0.1 of canopy height (z_1), respectively. Table III.7 in the Modelling and Mapping Manual shows the average relationship between O₃ concentrations at selected heights, derived from runs of the EMEP module over May-July, and selecting the noontime factors as representative of daytime multipliers. O₃ concentrations are normalised by setting the 20 m value to 1.0.

To use Table III.7, O₃ concentration measurements made above crops or grasslands may simply be extrapolated downwards to the canopy top for the respective vegetation. For example, with 30 ppb measured at 3 m height (above ground level) in a crop field, the concentration at 1 m would be $30.0 \text{ ppb} \times (0.88/0.95) = 27.8 \text{ ppb}$. For short grasslands we would obtain $30.0 \text{ ppb} \times (0.74/0.96) = 23.1 \text{ ppb}$ at canopy height 0.1 m. Experiments have shown that the vertical gradients found above for crops also apply well to tall (0.5m) grasslands. Some judgement may then be required to choose values appropriate to different vegetation types.

For forests, O₃ concentrations must often be derived from measurements made over grassy areas or other land cover types. In principle, the O₃ concentration measured over land-use X (e.g. short grasslands) could be used to estimate the O₃ concentration at a reference height, and then the gradient profile appropriate for desired land use Y could be applied. However, in order to keep this simple methodology manageable, and in view of the uncertainties inherent in making use of any profile near the canopy itself, it is suggested that concentrations are estimated by extrapolating the profiles given in Table III.7 upwards to the canopy height for forests. As an example, if we measure 30 ppb at 3 m above short grassland, the concentration at 20 m is estimated to be $30.0 \text{ ppb} \times (1.0/0.96) = 31.3 \text{ ppb}$.

It should be noted that the profiles shown in Table III.7 are representative only, and that site-specific calculations would provide somewhat different numbers. However, without local meteorology and the use of a deposition model, the suggested procedure should give an acceptable level of accuracy for most purposes. Gerosa et al. (2017.) verified that the tabulated gradients allowed them to calculate O₃ deposition fluxes for a mature forest that were in very good agreement with those directly measured with the eddy covariance technique.

Method b) Concentration profiles with stability effects

If all the important meteorological measurements are available, two possible cases can be envisaged: (1) if all the measurements were done above the desired (target) vegetation canopy or (2) if the measurements were done above a vegetated surface which is different from the desired surface.

Case 1: O₃ and meteorological measurements available above the target canopy

If the wind speed and the O₃ concentration were measured above the vegetation canopy, respectively at height $z_{m,w}$ and height z_{m,O_3} , the O₃ concentration at the top of the canopy (the target height z_{tgt}) can be obtained by making use of the big-leaf approximation and of the constant-flux assumption in the definition of the aerodynamic resistance. In the following, the roughness sub-layer affecting the concentration profiles near the canopy top has been neglected. A simple method for correcting for the roughness sub-layer can be found in Tuovinen and Simpson (2008). With the geometry illustrated in **Figure 3.2**, the O₃ concentration at the top of the canopy $O_3(z_{\text{tgt}})$ can be calculated with the following equation:

$$(1) \quad O_3(z_{tgt}) = O_3(z_{m,O_3}) \cdot \left[1 - \frac{R_a(z_{tgt}, z_{m,O_3})}{R_a(d + z_0, z_{m,O_3}) + R_b + R_{surf}} \right]$$

where $O_3(z_{m,O_3})$ is the available O_3 measurement above the canopy, $R_a(z_{tgt}, z_{m,O_3})$ is the aerodynamic resistance between the height where O_3 was measured and the top canopy height (the target height), $R_a(d + z_0, z_{m,O_3})$ is the aerodynamic resistance to O_3 deposition, i.e. the atmospheric resistance between the height where O_3 was measured and the height of the upper boundary of the laminar sub-layer of the theoretical big-leaf surface, R_b is the resistance to O_3 diffusion in the laminar sub-layer, and R_{surf} is the overall resistance to O_3 deposition to the underlying surfaces. The latter includes the stomatal resistance to O_3 uptake R_{stom} , the resistance of the external cuticles R_{ext} , the soil resistance to O_3 deposition R_{soil} and the air resistance to O_3 transfer within the vegetation layer R_{inc} .

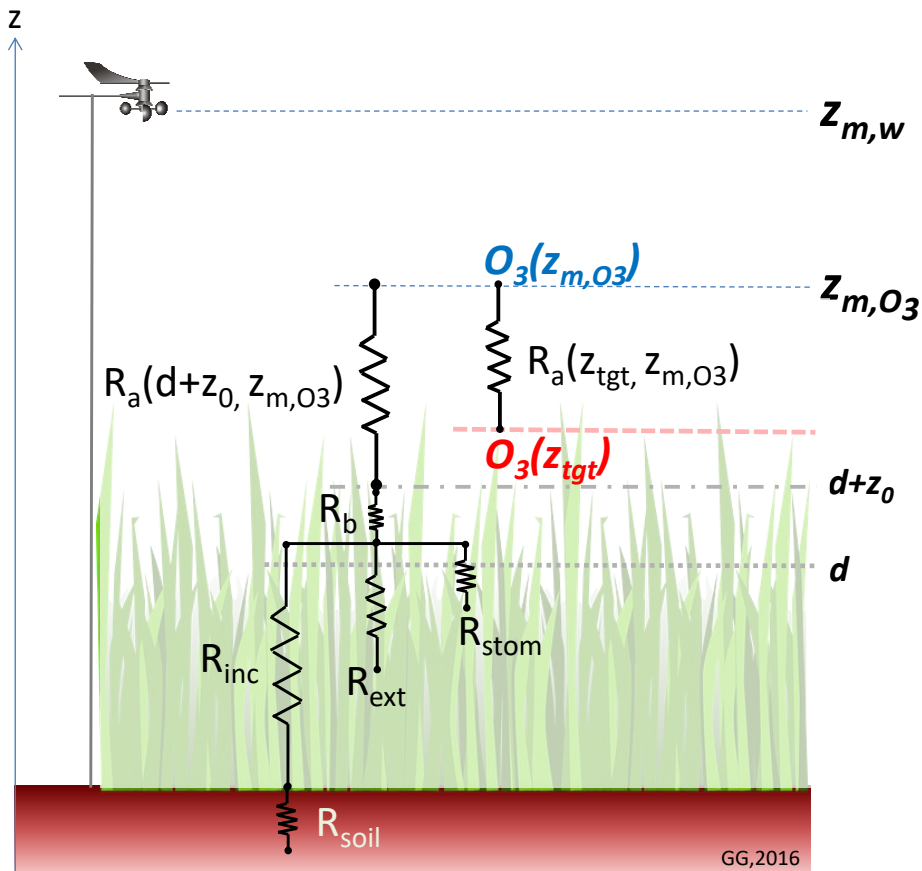


Figure 3.2 Resistive network for the calculation of the top-canopy O_3 concentration when O_3 and meteorological measurements are available ABOVE the vegetated surface (Case 1).

The calculation of the aerodynamic resistance R_a requires the estimation of the friction velocity u^* above the vegetated surface by the following equation

$$(2) \quad u^* = \frac{k \cdot u(z_{m,w})}{\ln\left(\frac{z_{m,w} - d}{z_0}\right) - \Psi_M\left(\frac{z_{m,w} - d}{L}\right) + \Psi_M\left(\frac{z_0}{L}\right)}$$

where k is the von Kármán constant (0.41), $u(z_{m,w})$ is the wind speed measured at the height $z_{m,w}$, d is the displacement height usually assumed as 2/3 of the canopy height, z_0 is the roughness length usually assumed as 1/10 of the canopy height, and $\Psi_M(\dots)$ is the

integral form of the similarity function for momentum which takes into account the stability of the atmospheric surface layer in terms of the Obukhov length L ($1/L = 0$ if the atmosphere is neutral, $1/L < 0$ if the atmosphere is unstable, $1/L > 0$ if the atmosphere is stable) (Garratt, 1992).

With ζ (adimensional length) as the argument, the function $\Psi_M(\zeta)$ is defined as (Dyer, 1974; Garratt, 1992):

$$(3) \quad \Psi_M(\zeta) = \begin{cases} \ln \left[\frac{1+x^2}{2} \cdot \left(\frac{1+x}{2} \right)^2 \right] - 2 \arctan(x) + \frac{\pi}{2}, & \text{when } \zeta < 0 \\ -5\zeta, & \text{when } \zeta \geq 0 \end{cases}$$

with $x = (1 - 16 \cdot \zeta)^{1/4}$.

Once u^* is known, the two R_a resistances in (1) can be calculated as follows:

$$(4) \quad R_a(z_{tgt}, z_{m,03}) = \frac{1}{k \cdot u^*} \left[\ln \left(\frac{z_{m,03} - d}{z_{tgt} - d} \right) - \Psi_H \left(\frac{z_{m,03} - d}{L} \right) + \Psi_H \left(\frac{z_{tgt} - d}{L} \right) \right]$$

$$(5) \quad R_a(d + z_0, z_{m,03}) = \frac{1}{k \cdot u^*} \left[\ln \left(\frac{z_{m,03} - d}{z_0} \right) - \Psi_H \left(\frac{z_{m,03} - d}{L} \right) + \Psi_H \left(\frac{z_0}{L} \right) \right]$$

with $\Psi_H(\dots)$, the similarity function for heat, defined as (Dyer, 1974; Garratt, 1992):

$$(6) \quad \Psi_H(\zeta) = \begin{cases} 2 \ln \left[\frac{1+x^2}{2} \right], & \text{when } \zeta < 0 \\ -5\zeta, & \text{when } \zeta \geq 0 \end{cases}$$

with $x = (1 - 16 \cdot \zeta)^{1/4}$.

The Obukhov length needed to account for the atmospheric stability can be estimated by following the procedure illustrated in the appendix of this chapter, or a guess value can be assumed according to the typical conditions of the area for which the fluxes should be calculated (e.g. $1/L=0$ for neutral conditions, $1/L=-0.01$ for unstable or $1/L=-0.1$ for very unstable conditions, $1/L=+0.01$ for stable conditions). If no information on atmospheric stability is known at all, the $\Psi_M(\dots)$ and $\Psi_H(\dots)$ functions in (1), (4) and (5) can be set to zero and then the atmosphere is assumed as neutral.

The resistance to O_3 diffusion in the laminar sub-layer R_b can be calculated with the formulation of Wesely and Hicks (1977):

$$(7) \quad R_b = \frac{2}{k \cdot u^*} \left(\frac{Sc}{Pr} \right)^{2/3}$$

where k is the von Kármán constant, $Sc=0.93$ is the Schmidt number for O_3 , and $Pr=0.71$ is the Prandtl number of air.

The surface resistance to O_3 deposition R_{surf} is defined as follows:

$$(8) \quad R_{surf} = \frac{1}{\frac{LAI}{R_{sto}} + \frac{SAI}{R_{ext}} + \frac{1}{R_{inc} + R_{soil}}}$$

where R_{sto} is the leaf-scale stomatal resistance to O_3 of the vegetated surface, R_{ext} is the leaf-scale resistance of the external vegetation surfaces (e.g. cuticles, bark, etc.) to O_3 deposition, R_{soil} is the soil resistance to O_3 deposition, R_{inc} is the in-canopy air resistance to the O_3 transfer to the soil, LAI is the projected leaf area ($m^2 m^{-2}$), and SAI is the surface area of the canopy (green LAI + senescent LAI + twig and branch surfaces).

R_{sto} is a plant species specific function of air temperature and humidity, solar radiation and soil water content. It can be modelled – as described in the manual – by means of the Jarvis–Stewart algorithm (Jarvis, 1976; Emberson et al., 2000a) and data on the photosynthetically active radiation (light), air temperature (temp) and relative humidity (to calculate vapour pressure deficit - VPD) at the canopy top (Tuovinen and Simpson, 2008), the soil water potential (SW), and plant phenology (phen):

$$(9) \quad R_{sto} = \frac{1}{g_{sto}} = \frac{1}{g_{max} \cdot [\min(f_{phen}, f_{O3})] \cdot f_{light} \cdot \max\{f_{min}, (f_{temp} \cdot f_{VPD} \cdot f_{SW})\}}$$

where g_{max} is the stomatal conductance of O_3 in non-limiting conditions and the f functions f_{phen} , f_{light} , f_{temp} , f_{VPD} , f_{SW} are species-specific functions which describe the variation of the stomatal conductance with phenology, light, air temperature, leaf-to-air vapour pressure deficit and soil water potential, respectively. For details, see Section III.3.4.3 of the manual on the modelling of the stomatal conductance.

The resistance to cuticular deposition of O_3 , R_{ext} , and the soil resistance, R_{soil} , are set respectively to $2500 s m^{-1}$ and $200 s m^{-1}$ for consistency with the EMEP model (Simpson et al., 2012).

The in-canopy resistance R_{inc} can be calculated according to van Pul & Jacobs (1994):

$$(10) \quad R_{inc} = b \cdot SAI \cdot h/u^*$$

where $b = 14 m^{-1}$ is an empirical constant, h is the height of the canopy and SAI is the surface area of the canopy.

It is worth noticing that all the resistances in the above equations are expressed in the unit of $s m^{-1}$. Thus, the stomatal conductance, usually given in $mmol m^{-2} s^{-1}$, should be converted to $m s^{-1}$. The conversion can be done by multiplying g_{sto} by $R \cdot T/P$ with the gas constant $R = 8.314 J mol^{-1} K^{-1}$, the air temperature T in Kelvin and the atmospheric pressure P in Pa. It should be noted that g_{stom} is the stomatal conductance to O_3 , and not to water vapour; stomatal conductance to water vapour can be converted to O_3 by multiplying it by 0.663.

Case 2: Calculation of the O_3 concentration over a target surface from measurements made above a different surface

This case typically occurs with forests, for which O_3 concentrations must be derived from measurements made over grassy areas or other land-cover types (Gerosa et al., 2017; Tuovinen et al., 2009). The O_3 concentrations measured over e.g. short grasslands are used to estimate the O_3 concentrations at a reference height that is greater than the forest height, and then the gradient profile appropriate for the forest surface is applied to derive the concentrations at the top of the forest canopy. It should be noted, however, that there are potentially significant uncertainties involved in this approach. In addition to the roughness sub-layer effect mentioned above (Tuovinen & Simpson, 2008), the application of the profile functions presented here inherently assume extensive homogeneous surfaces (i.e. an adequate fetch). The validity of the method is compromised if the measurements are taken close to the forest edge, see for example Tuovinen et al. (2009).

The typical situation is that described in **Figure 3.3**, where the O_3 concentration $O_3(z_{m,O_3})$ and the wind speed $u(z_{m,O_3})$ are measured over a reference surface (e.g. grassland) at a measuring height z_{m,O_3} and $z_{m,w}$ respectively. The aerodynamic features of the reference surface (h , d , z_0 , LAI, SAI) as well as all the resistances to O_3 deposition (r_a , r_b , r_{stom} , r_{ext} , r_{inc} , r_{soil}) over the reference surface should be known. The same is true for the target surface (e.g. forest), for which h , d , z_0 , LAI, SAI, as well as all the resistances to O_3 deposition (R_a , R_b , R_{stom} , R_{ext} , R_{inc} , R_{soil}) should be known.

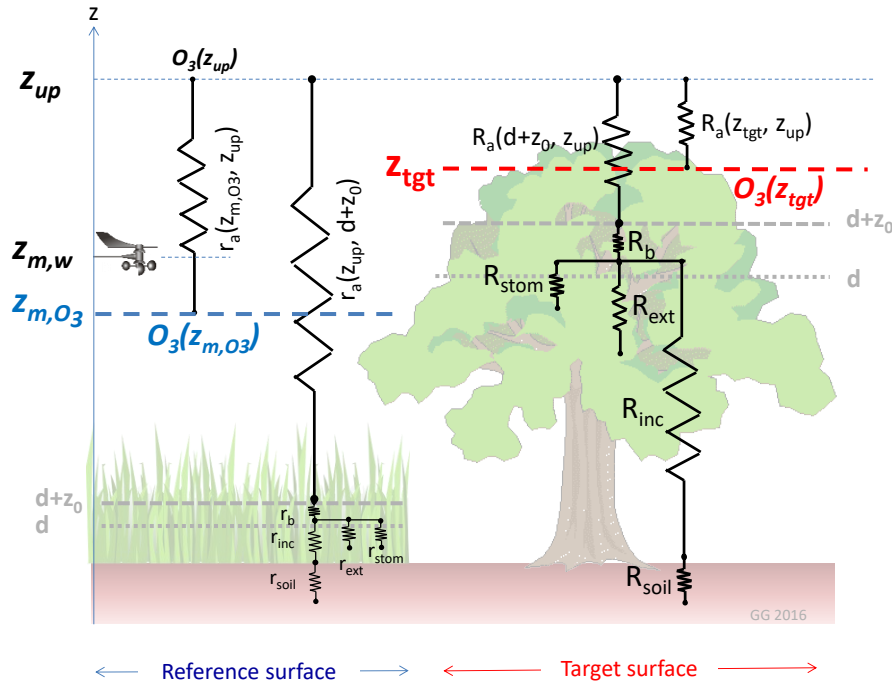


Figure 3.3 Resistive network for the calculation of the O_3 concentration at the top of a target canopy (e.g. forest) when O_3 and meteorological measurements are available above a different vegetated surface (e.g. grassland) (Case 2).

The estimation of the O_3 concentration at the top of the target canopy $O_3(z_{tgt})$ requires two steps:

- i) the calculation of the O_3 concentration at an height (above the reference and the target surfaces) where it is not influenced by variations in the underlying surface;
- ii) the calculation of the O_3 concentration at the desired height above the target surface.

Step i) Calculation of the O_3 concentration at a height where it is not influenced by surface variability:

First of all the O_3 concentration should be calculated at a height z_{up} at which it is not influenced by variation in the properties of the underlying surface. This height is usually assumed at 50 m (Simpson et al., 2012).

For this sake the friction velocity u^* above the reference surface should be calculated analogous to (2):

$$(11) \quad u_{ref}^* = \frac{k \cdot u(z_{m,w})}{\ln\left(\frac{z_{m,w} - d_{ref}}{z_{0,ref}}\right) - \Psi_M\left(\frac{z_{m,w} - d_{ref}}{L}\right) + \Psi_M\left(\frac{z_{0,ref}}{L}\right)}$$

where the subscripts ref indicate that the parameters refer to the reference surface (e.g. grassland). The Obukhov length and the $\Psi_M(\dots)$ function have been already introduced in the previous section.

Then the O_3 concentration at the height z_{up} is given by:

$$(12) \quad O_3(z_{up}) = \frac{O_3(z_{m,03})}{1 - \frac{r_a(z_{m,03}, z_{up})}{r_a(d_{ref} + z_0, z_{up}) + r_b + r_{surf}}}$$

where the two atmospheric resistances are given by the following expressions:

$$(13) \quad r_a(z_{m,03}, z_{up}) = \frac{1}{k \cdot u_{ref}^*} \left[\ln \left(\frac{z_{up} - d_{ref}}{z_{m,03} - d_{ref}} \right) - \Psi_H \left(\frac{z_{up} - d_{ref}}{L} \right) + \Psi_H \left(\frac{z_{m,03} - d_{ref}}{L} \right) \right]$$

$$(14) \quad r_a(d_{ref} + z_0, z_{up}) = \frac{1}{k \cdot u_{ref}^*} \left[\ln \left(\frac{z_{up} - d_{ref}}{z_0,ref} \right) - \Psi_H \left(\frac{z_{up} - d_{ref}}{L} \right) + \Psi_H \left(\frac{z_0,ref}{L} \right) \right]$$

with the friction velocity u_{ref}^* calculated with (11) and the similarity function $\Psi_H(\dots)$ defined as in (6). The resistance r_b is calculated with (7) by using u_{ref}^* for u^* . The resistance r_{surf} is calculated in a way analogous to what was explained for the case 1 ((8), (9) and (10)) by taking into account the appropriate geometry, LAI, SAI and the f functions for the reference canopy, and by setting r_{ext} and r_{soil} to the values indicated for R_{ext} and R_{soil} above.

Step ii) Calculation of the O_3 concentration at the desired height above the target surface:

Once the O_3 concentration at the height z_{up} is known, the O_3 concentration at the *target* height canopy $O_3(z_{tgt})$ above the *target* surface can be calculated.

First the friction velocity above the target surface u_{tgt}^* should be calculated as:

$$(15) \quad u_{tgt}^* = \frac{k \cdot u(z_{up})}{\ln \left(\frac{z_{up} - d_{tgt}}{z_0,tgt} \right) - \Psi_M \left(\frac{z_{up} - d_{tgt}}{L} \right) + \Psi_M \left(\frac{z_0,tgt}{L} \right)}$$

where d and z_0 now refer to the target surface ('tgt' suffix). The wind speed at the height z_{up} which appears in (15) as $u(z_{up})$ – the height at which the wind is assumed not to be influenced by variations in the underlying surface – can be calculated by the following formula:

$$\text{Equation (16)} \quad u(z_{up}) = \frac{u_{ref}^*}{k} \cdot \left[\ln \left(\frac{z_{up} - d_{ref}}{z_0,ref} \right) - \Psi_M \left(\frac{z_{up} - d_{ref}}{L} \right) + \Psi_M \left(\frac{z_0,ref}{L} \right) \right]$$

where here d and z_0 refer to the reference surface (i.e. the grassland).

Then the O_3 concentration at the desired height z_{tgt} above the target surface is given by:

$$(17) \quad O_3(z_{tgt}) = O_3(z_{up}) \cdot \left[1 - \frac{R_a(z_{tgt}, z_{up})}{R_a(d + z_0, z_{up}) + R_b + R_{surf}} \right]$$

where $O_3(z_{up})$ was calculated with (12) and the two atmospheric resistances – which refer to the target surface – are given by the following expressions:

$$(18) \quad R_a(z_{tgt}, z_{up}) = \frac{1}{k \cdot u_{tgt}^*} \left[\ln \left(\frac{z_{up} - d_{tgt}}{z_{tgt} - d_{tgt}} \right) - \Psi_H \left(\frac{z_{up} - d_{tgt}}{L} \right) + \Psi_H \left(\frac{z_{tgt} - d_{tgt}}{L} \right) \right]$$

$$(19) \quad R_a(d + z_0, z_{up}) = \frac{1}{k \cdot u_{tgt}^*} \left[\ln \left(\frac{z_{up} - d_{tgt}}{z_{0,tgt}} \right) - \Psi_H \left(\frac{z_{up} - d_{tgt}}{L} \right) + \Psi_H \left(\frac{z_{0,tgt}}{L} \right) \right]$$

with the friction velocity u_{tgt}^* calculated with (15) and the similarity function $\Psi_H(\cdot)$ defined as in (6). The resistance R_b is calculated with (7) by using u_{tgt}^* for u^* . The resistance R_{surf} is calculated in a way analogous to what was explained for Case 1 ((8), (9) and (10)) by taking into account the appropriate geometry, LAI, SAI and the f functions for the target canopy. A comparison among the O_3 concentrations measured at the top of the forest and those estimated at the same level with this methodology, by using measurements taken on a nearby grassland, can be found in Gerosa et al. (2017). The paper discusses also the uncertainties related to different calculation options and their consequences on the estimated POD_1SPEC in comparison to the ozone fluxes measured by eddy covariance at the same site.

Method c) Concentration profiles with no stability correction

If no information on atmospheric stability is available, method b) can be simplified by ignoring the stability correction by setting the value of $\Psi_M(\cdot)$ and $\Psi_H(\cdot)$ functions in (1), (4) and (5) (the shaded terms) to zero. By doing so, the atmospheric stability is assumed to be neutral.

3.3 References

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3.4 Appendix

Estimation of the Obukhov length (L)

The Obukhov length L (m) is an indicator of the atmospheric stability, but its calculation requires that some other parameters are estimated aside. L is defined by the following equation:

$$(20) \quad L = - \frac{u_*^3}{k \frac{g}{T} \frac{H}{\rho c_p}}$$

where u^* is the friction velocity (m/s), k is the Von Kármán constant (0.41, adim), g is the gravity acceleration (9.8 m s^{-2}), T is the air temperature (K), H is the sensible heat flux (W m^{-2}), ρ is the air density (kg m^{-3}), c_p is the specific heat at constant pressure ($1048 \text{ J kg}^{-1} \text{ K}^{-1}$). Not all these data are usually available from traditional slow meteorological stations, in particular u^* and H . Relatively easy measurements of u^* and H can be performed with an ultrasonic anemometer but nearly always it is not available. Hence, to estimate L a model of H and u^* , and also of the net radiation (R_n) which is required for the H estimation, are needed.

Estimation of the net radiation (Rn)

Net radiation can be estimated using the methodology proposed by Holtslag & Van Ulden (1983):

$$(21) \quad R_n = \frac{((1 - A)Q_{sw} + c_1 T^6 - \sigma T^4 + c_2 N)}{1 + c_3}$$

where A is the albedo (fraction between 0..1), T is the air temperature (K), N is the cloud cover (%), c_1 and c_2 are constants (whose values are respectively $5.31 \cdot 10^{-13} \text{ W m}^{-2}$ and 60 W m^{-2}), σ is the Stefan-Boltzmann constant ($5.67 \text{ E-}08 \text{ W m}^{-2} \text{ K}^{-4}$), Q_{sw} is the shortwave radiation (the global radiation which is typically available from traditional meteorological stations, W m^{-2}) and c_3 is a temperature dependent parameter which will be presented few lines below.

The cloud cover N can be estimated from the measured shortwave radiation taking into account the solar elevation angle (ν , degree) with the following equation taken from Holtslag & Van Ulden (1983):

$$(22) \quad N = \sqrt[3]{\frac{b_2}{b_1} \left(1 - \frac{Q_{SW}}{(990 \sin v - 30)} \right)}$$

where: b_1 and b_2 are empirical constants whose values are respectively 0.75 and 3.4. The solar elevation angle v (degrees) can be calculated by downloading the tool available in the NOAA web site (<http://www.esrl.noaa.gov/gmd/grad/solcalc/calcdetails.html>). The c_3 parameter is obtained by the following equation:

$$(23) \quad c_3 = 0.38 \cdot ((1 - \alpha) \cdot S + 1) / (S + 1)$$

where α is the water availability parameter described in Beljaars and Holtslag (1989 and 1991) and whose values can be taken from **Table A3.1** (Hanna & Chang, 1992), and S is a temperature dependent parameter described by the following equation derived from the tabulated values of Hanna & Chang (1992):

$$(24) \quad S = 1.5 \cdot e^{-0.060208041 \cdot T}$$

with T the air temperature in Celsius degrees.

Table A3.1 Values for the parameter α proposed by Hanna & Chang (1992).

Values for the parameter α		
From	To	Description
0	0.2	Arid desert without rainfalls for months
0.2	0.4	Rural arid area
0.4	0.6	Agricultural fields in periods with no rainfalls for long periods
0.5	1	Urban environment
0.8	1.2	Agricultural fields or forests with sufficient water availability
1.2	1.4	Big lake or ocean, far at least 10 km from the shore

Estimation of the sensible heat flux (H)

Sensible heat fluxes can be modelled using the methodology proposed by Holtslag & Van Ulden (1983):

$$(25) \quad H = \frac{(1 - \alpha) + S}{1 + S} (R_n + Q_A - G) - \alpha \beta$$

where Q_A is the anthropogenic heat flux (which is always set equal to zero as suggested by Hanna & Chang, 1992), S and α are respectively the temperature dependent parameter and the water availability parameter just described above ((24) and Table A3.1), β is a constant value equal to 20 W m^{-2} which takes into account that sensible heat flux is usually negative just before the sunset (Hanna & Chang 1992), and G is the ground heat flux assumed as a fraction of the net radiation

$$(26) \quad G = a \cdot R_n$$

with a constant value ($a=0.1$ for rural areas and $a=0.3$ for urban areas) taken from Doll et al. (1985). During the nighttime hours ($R_n < 50 \text{ W m}^{-2}$) the sensible heat flux is calculated as $H = -\alpha\beta$.

Estimation of the friction velocity (u^*)

The friction velocity can be estimated following the methodology proposed in Bassin et al. (2003). When $H < 1 \text{ W m}^{-2}$ (Stable atmosphere) u^* is calculated by the following equation:

$$(27) \quad u^* = \frac{0.5 k \cdot U}{\ln((z_{ref} - d)/z_0)} \left(1 + \sqrt{1 - \frac{4(5 \cdot g \cdot z_{ref} \cdot \theta^* \cdot \ln((z_{ref} - d)/z_0))}{k \cdot T_0 \cdot U^2}} \right)$$

where k is the Von Kármán constant, U is the horizontal wind speed (m s^{-1}), z_{ref} is the measurement height of the wind speed (m), d is the displacement height (m) usually taken as $2/3$ of the canopy height, T_0 are the Kelvin degrees at 0°C (i.e. $T_0=273.15 \text{ K}$), z_0 is the roughness length (m) (z_0 values can be taken from the table at page 1.5-12 of WMO, 2008), g is the gravity acceleration (m s^{-2}) and θ^* is the scale temperature (K) calculated according to the following equation:

$$(28) \quad \theta^* = \frac{-H}{\rho \cdot c_p \cdot u_{neutral}^*}$$

where $u_{neutral}^*$ (m s^{-1}) with the following equation:

$$(29) \quad u_{neutral}^* = \frac{U}{k \ln((z_{ref} - d)/z_0)}$$

When $H > 1 \text{ W m}^{-2}$ (unstable atmosphere, the friction velocity is calculated by the following equation:

$$(30) \quad u^* = \frac{k U}{\ln[(z_{ref} - d)/z_0]} [1 + d_1 \ln(1 + d_2 d_3)]$$

where d_1 , d_2 and d_3 are respectively:

$$(31) \quad d_1 = \begin{cases} 0.128 + 0.005 \ln \left[\frac{z_0}{(z_{ref} - d)} \right] & \text{if } \frac{z_0}{z_{ref} - d} \leq 0.01 \\ 0.107 & \text{if } \frac{z_0}{z_{ref} - d} > 0.01 \end{cases}$$

$$(32) \quad d_2 = 1.95 + 32.6 \left(\frac{z_0}{z_{ref} - d} \right)^{0.45}$$

$$(33) \quad d_3 = \frac{H}{\rho c_p} \frac{k g (z_{ref} - d)}{T_0} \left(\frac{\ln((z_{ref} - d)/z_0)}{k U} \right)^3$$

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4 Crops – Additional information on O₃ flux model parameterisation (Section III.3.5.2 of the manual)

4.1 Wheat and potato

The g_{\max} values for wheat (*Triticum aestivum*) and potato (*Solanum tuberosum*) for Atlantic, Boreal and Continental regions have been derived from published data conforming to a strict set of criteria for use in establishing this key parameter of the flux algorithm. Only data obtained from g_{sto} measurements made on cultivars grown either under field conditions or using field-grown plants in open top chambers in Europe were considered. Measurements had to be made during those times of the day and year when g_{\max} would be expected to occur and full details had to be given of the gas for which conductance measurements were made (e.g. H₂O, CO₂, O₃) and the leaf surface area basis on which the measurements were given (e.g. total or projected). All g_{sto} measurements were made on the flag leaf for wheat and for sunlit leaves of the upper canopy for potato using recognized g_{sto} measurement apparatus. **Tables 4.1 and 4.2** give details of the published data used for g_{\max} derivation on adherence to these rigorous criteria. **Figure 4.1** shows the mean, median and range of g_{\max} values for each of the 14 and four different cultivars that provide the approximated g_{\max} values of 500 and 750 mmol O₃ m⁻² PLA s⁻¹ for wheat and potato, respectively.

It should be noted that the wheat g_{\max} value has been parameterised from data collected for spring and winter wheat cultivars. For potato additional g_{\max} values from three USA grown cultivars are included in Figure 4.1 for comparison (Stark, 1987), further substantiating the g_{\max} value established for this crop type.

González-Fernández et al. (2013) compiled data from 25 years of phenology data from areas representative of the Mediterranean region with Atlantic climate influence, coastal Mediterranean and continental Mediterranean climates in Spain together with stomatal conductance measurements made over five years for winter bread wheat (3 cultivars) and durum wheat (2 cultivars) growing near Madrid. In this study, g_{\max} was derived from a literature review of wheat growing under Mediterranean conditions (10 cultivars of bread wheat and 2 cultivars of durum wheat). For further details including boundary line plots for the component parameters, see González-Fernández et al. (2013). Parameterisations of the O₃ stomatal flux model are provided in Table III.9 of the manual.

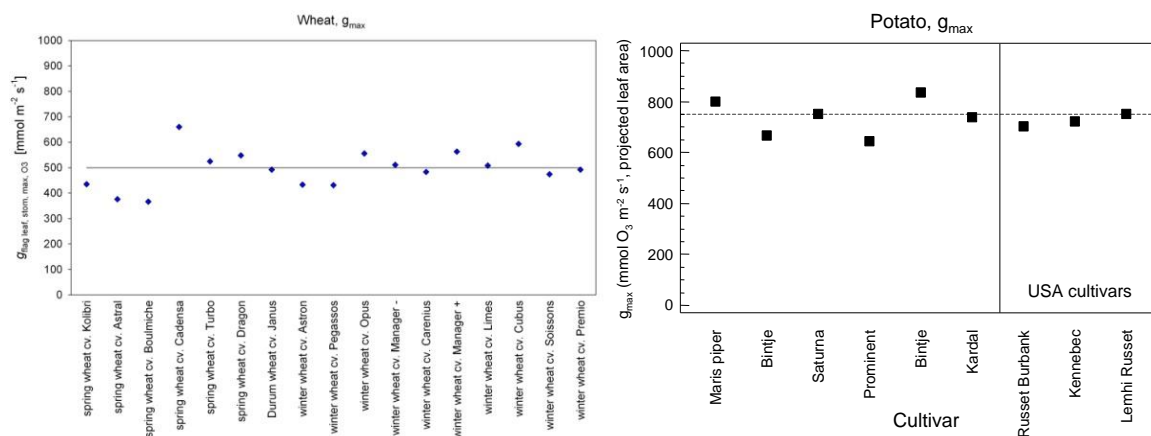


Figure 4.1 Derivation of g_{\max} for wheat and potato (see Tables 4.1 and 4.2 for details).

Table 4.1 Derivation of wheat g_{max} parameterisation. PLA = projected leaf area. The data used was first published in Pleijel et al. (2007) and updated in Grünhage et al. (2012).

Reference	g_{max} [mmol O ₃ m ⁻² s ⁻¹ PLA]	g_{max} derivation	Country	Wheat type and cultivar	Time of day	Time of year	g_{sto} measuring apparatus	Gas / leaf area	Growing conditions	Leaf
Araus et al. (1989)	435	Value in Table. Cultivar and sowing time (average of 3) g_{sto} used. Means of 5 to 7 replicates. g_{sto} mmol CO ₂ m ⁻² s ⁻¹ . Adaxial: 313, abaxial: 149	Spain	Spring wheat, Kolibri	9 to 13 hrs	14 March to 21 May	LI 1600 steady state porometer	CO ₂ / PLA	Field	Flag
Araus et al. (1989)	376	Value in Table. Means and SE ± of 5 to 7 replicates. g_{sto} mmol CO ₂ m ⁻² s ⁻¹ . Adaxial: 267 ± 29, abaxial: 92 ± 16.	Spain	Spring wheat, Astral	9 to 13 hrs	14 March to 21 May	LI 1600 steady state porometer	CO ₂ / PLA	Field	Flag
Araus et al. (1989)	366	Value in Table. Means and SE ± of 5 to 7 replicates. g_{sto} mmol CO ₂ m ⁻² s ⁻¹ . Adaxial: 251 ± 15, abaxial: 99 ± 22.	Spain	Spring wheat, Boulmiche	9 to 13 hrs	14 March to 21 May	LI 1600 steady state porometer	CO ₂ / PLA	Field	Flag
Ali et al. (1999)	660	From graph showing leaf conductance plotted against time in days. Maximum approximately 1 mol H ₂ O m ⁻² s ⁻¹ ; ± 0.12. ± SE of 4 to 6 replicates.	Denmark	Spring wheat, Cadensa	(Assumed mid-day)	August	IRGA LI-6200	H ₂ O / * (Assume PLA as use LAI)	Field Lysimeter	Flag
Grüters et al. (1995)	525	Value in text. Maximum measured conductance (0.97 cm s ⁻¹ H ₂ O total leaf area after Jones (1983)).	Germany	Spring wheat, Turbo	11 to 12 hrs	17 June to 7 August	LI 1600 steady state porometer	H ₂ O / total leaf area	Field	Flag
Danielsson et al. (2003)	548	Value in text. "The maximum conductance value, 414 mmol H ₂ O m ⁻² s ⁻¹ , was taken as g_{max} for the Östad multiplicative model. The conductance values represent the flag leaf and are given per total leaf area".	Sweden	Spring wheat, Dragon	13 hrs	13 August 1996 (AA)	Li-Cor 6200	H ₂ O / total leaf area	Field OTC & AA	Flag
Körmer et al. (1979)	492	Value given in table. 0.91 cm s ⁻¹ for H ₂ O on a total leaf surface area basis.	Austria	Durum wheat, Janus	-	-	Ventilated diffusion porometer	H ₂ O / total leaf area	Field	Flag
Grünhage et al. (2012)	433	653 mmol H ₂ O m ⁻² s ⁻¹	Germany	Winter wheat, Astron	measured at 10 hrs	24 May to 14 June 2006	Li-Cor 6400	H ₂ O / total leaf area	OTC (NF)	Flag
Grünhage et al. (2012)	431	650 mmol H ₂ O m ⁻² s ⁻¹	Germany	Winter wheat, Pegassos	measured at 10 CET	24 May to 14 June 2006	Li-Cor 6400	H ₂ O / total leaf area	OTC (NF)	Flag
Grünhage et al. (2012)	556	839 mmol H ₂ O m ⁻² s ⁻¹ (adaxial=524, abaxial=315)	Germany	Winter wheat, Opus	measured at 11 CET	26 May to 02 June 2009	Leaf porometer SC-1	H ₂ O / PLA	Field	Flag
Grünhage et al. (2012)	511	770 mmol H ₂ O m ⁻² s ⁻¹ (adaxial=439, abaxial=331)	Germany	Winter wheat, Manager -	measured at 10 CET	26 May to 02 June 2009	Leaf porometer SC-1	H ₂ O / PLA	Field	Flag
Grünhage et al. (2012)	483	729 mmol H ₂ O m ⁻² s ⁻¹ (adaxial=451, abaxial=278)	Germany	Winter wheat, Carenius	measured at 13 CET	26 May to 02 June 2009	Leaf porometer SC-1	H ₂ O / PLA	Field	Flag
Grünhage et al. (2012)	563	849 mmol H ₂ O m ⁻² s ⁻¹ (adaxial=485, abaxial=364)	Germany	Winter wheat, Manager +	measured at 11:30 CET	26 May to 02 June 2009	Leaf porometer SC-1	H ₂ O / PLA	Field	Flag
Grünhage et al. (2012)	508	766 mmol H ₂ O m ⁻² s ⁻¹ (adaxial=510, abaxial=256)	Germany	Winter wheat, Limes	measured at 11:30 CET	26 May to 02 June 2009	Leaf porometer SC-1	H ₂ O / PLA	Field	Flag
Grünhage et al. (2012)	593	894 mmol H ₂ O m ⁻² s ⁻¹ (adaxial=595, abaxial=299)	Germany	Winter wheat, Cubus	measured at 11:30 CET	20 May to 02 June 2009	Leaf porometer SC-1	H ₂ O / PLA	Field	Flag
Grünhage et al. (2012)	474	714.4 ± 42.1 mmol H ₂ O m ⁻² s ⁻¹	France	Winter wheat, Soissons	11 to 16 CET	6 to 27 May 2009	PP systems CIRAS-2	H ₂ O / PLA	Field	Flag
Grünhage et al. (2012)	492	741.6 ± 72.8 mmol H ₂ O m ⁻² s ⁻¹	France	Winter wheat, Premio	11 to 16 CET	6 to 27 May 2009	PP systems CIRAS-2	H ₂ O / PLA	Field	Flag
Mean :	497									
Median :	492	Range: 366 to 660 mmol O ₃ m ⁻² s ⁻¹								

Table 4.2 Derivation of potato g_{max} parameterisation. PLA = projected leaf area.

Reference	g_{max} [mmol O ₃ m ⁻² s ⁻¹ PLA]	g_{max} derivation	Country	Potato cultivar	Time of day	Time of year	g_{sto} measuring apparatus	Gas / leaf area	Growing conditions	Leaf
Jeffries (1994)	800	Value given in Figure. Maximum value of 16 mm s ⁻¹ . Error bar represents SE of the difference between two means (n=48).	Scotland	Maris piper	8 to 16 hrs	June	Diffusion porometer	Assumed H ₂ O / assumed PLA	Field	Fully expanded in upper canopy
Vos & Groenwald (1989)	665	Value given in Figure. Maximum value of 13.3 mm s ⁻¹ . Replicates approx. 20, the coefficient of variation typically ranged from 15 to 25%.	Netherlands	Bintje	-	June / July	Li-Cor 1600 steady state diffusion porometer	H ₂ O / PLA	Field	Youngest fully grown leaf
Vos & Groenwald (1989)	750	Value given in Figure. Maximum value of 15 mm s ⁻¹ . Replicates approx. 20, the coefficient of variation typically ranged from 15 to 25%.	Netherlands	Saturna	-	June	Li-Cor 1600 steady state diffusion porometer	H ₂ O / PLA	Field	Youngest fully grown leaf
Marshall & Vos (1991)	643	Value given in Figure. g_{max} of 527 mmol H ₂ O m ⁻² s ⁻¹ at intermediate N supply. Each point represents the mean of at least three leaves (usually four).	Netherlands	Prominent	-	July	LCA2 portable infra-red gas analyser	H ₂ O / assumed PLA	Field	Most recently expanded measurable leaf
Pleijel <i>et al.</i> (2002)	836	Value given in Table. g_{max} of 1371 mmol m ⁻² s ⁻¹ for H ₂ O per projected leaf area.	Germany	Bintje	12	June	Li-Cor 6200	H ₂ O / PLA	Field	Fully expanded in upper canopy
Danielsson (2003)	737	Value given in text. g_{max} of 604 mmol H ₂ O m ⁻² s ⁻¹ per total leaf area.	Sweden	Kardal	11	July	Li-Cor 6200	H ₂ O / Total leaf area	Field	Fully expanded in upper canopy
Mean	738	Range: 643 to 836								
Median	743									

f_{min}

The data presented in Pleijel et al. (2002) and Danielsson et al. (2003) clearly show that for both species, f_{min} under field conditions frequently reaches values as low as 1% of g_{max} . Hence an f_{min} of 1% of g_{max} is used to parameterise the model for both species.

f_{phen}

The data used to establish the f_{phen} relationships for both wheat and potato are given in **Figure 4.2** as °C days from g_{max} (in the case of wheat g_{max} is assumed to occur between growth stages "flag leaf fully unrolled, ligule just visible" and "mid-anthesis"; in the case of potato g_{max} is assumed to occur at the emergence of the first generation of fully developed leaves). Methods for estimating the timing of mid-anthesis and for estimating f_{phen} using the functions illustrated in Figure 4.2 are provided in Section III.3.5.2.1, with the parameterisations given in Table III.9 of the manual.

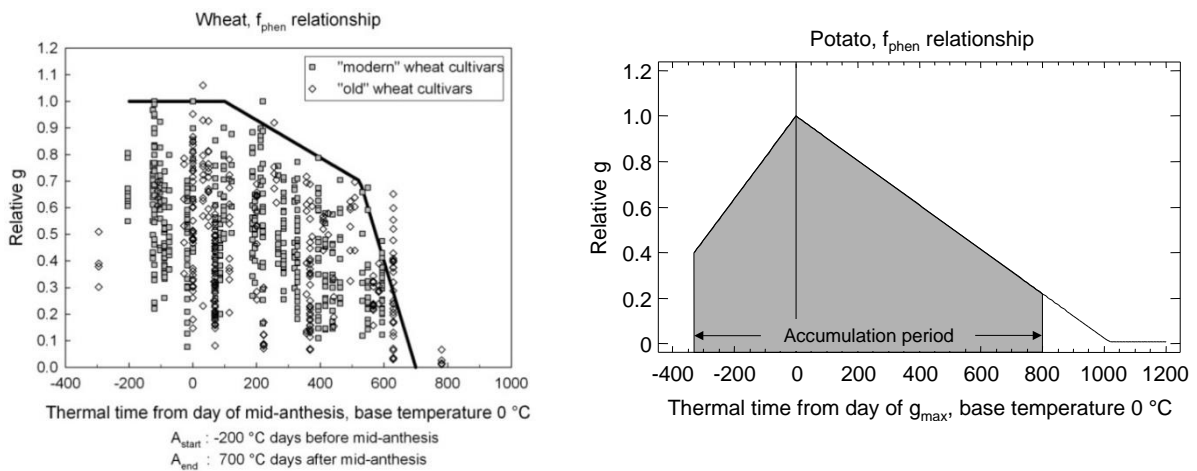


Figure 4.2 f_{phen} functions for (a) wheat and (b) potato. The potato function was published in Pleijel et al., 2007; the wheat function has since been revised, with new data from Grünhage et al. (2012).

f_{light}

The data used to establish the f_{light} relationship for both wheat and potato are shown in Figure 4.3.

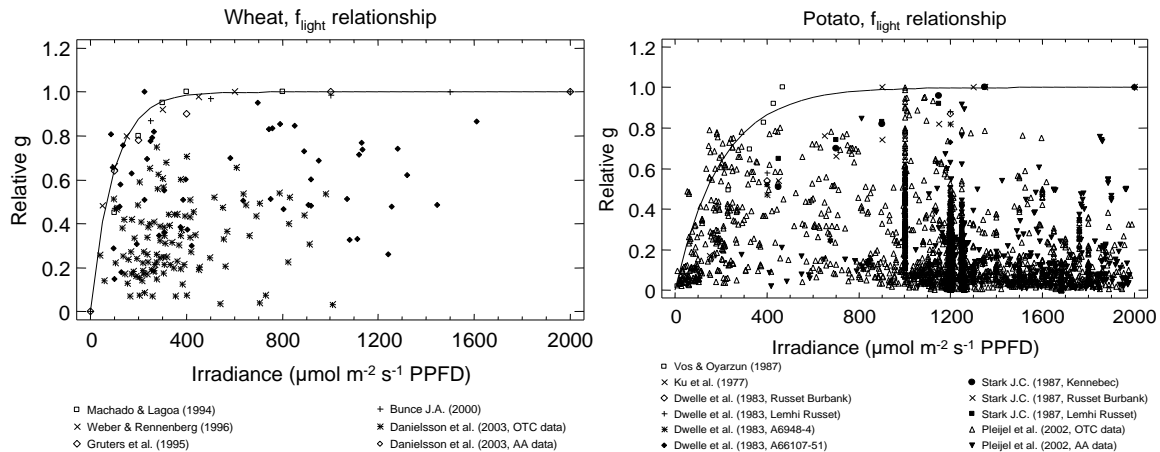


Figure 4.3 Derivation of f_{light} for wheat and potato (see Pleijel et al., 2007 for further details).

f_{temp}

The data used to establish the f_{temp} relationship for both wheat and potato are shown in **Figure 4.4**.

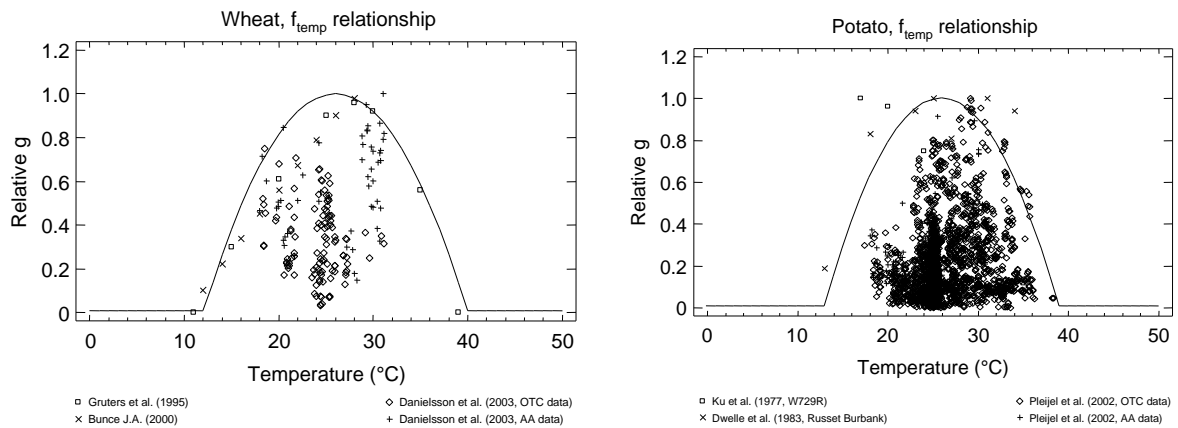


Figure 4.4 Derivation of f_{temp} for wheat and potato (see Pleijel et al., 2007 for further details).

f_{VPD} and ΣVPD_{crit}

The data used to establish the f_{VPD} relationship for both wheat and potato are shown in **Figure 4.5**. Under Mediterranean conditions, an alternative parameterization for VPD is provided that has been derived from **Figure 4.6**. Values of ΣVPD_{crit} for wheat and potato are given in Table III.9 of the manual.

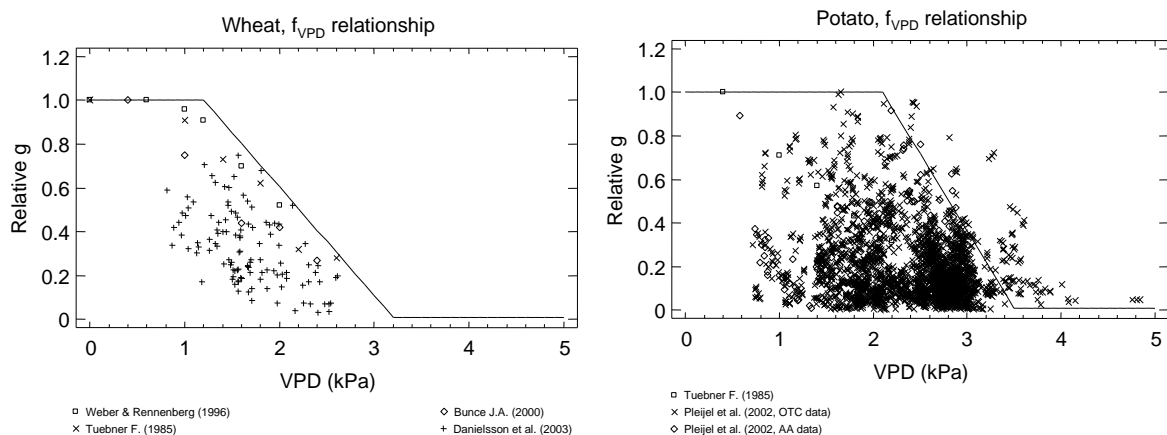


Figure 4.5 Derivation of f_{VPD} for wheat and potato (see Pleijel et al., 2007 for further details).

f_{PAW} for wheat and f_{SWP} for potato

The shape of the f_{PAW} function for wheat and the f_{SWP} for potato are shown in Figure III.9 and III.8 of the manual respectively. It should be noted that the f_{SWP} relationship for potato is derived from data that describe the response of potato g_{sto} to leaf water potential rather than soil water potential. Vos and Oyarzun (1987) state that their results represent long-term effects of drought, caused by limiting supply of water rather than by high evaporative demand, and hence can be assumed to apply to situations where pre-dawn leaf water potential is less than 0.1 to 0.2 MPa. As such, it may be necessary to revise this f_{SWP} relationship so that potato g_{sto} responds more sensitively to increased soil water stress.

4.2 Tomato

The parameterisation for tomato (*Solanum lycopersicum*) was derived from g_{sto} measurements made on seven cultivars grown in pots under open-top chambers conditions in southern Europe (Spain and Italy; González-Fernández et al., 2014). Daily profiles of g_{sto} were measured in different days from July to October, under varying environmental conditions. All g_{sto} measurements were made in sunlit leaves of the upper canopy using standard g_{sto} measurement systems. The g_{max} value for tomato was set as the average of the values above the 95th percentile of all the g_{sto} measurements (**Figure 4.6, Table 4.3**).

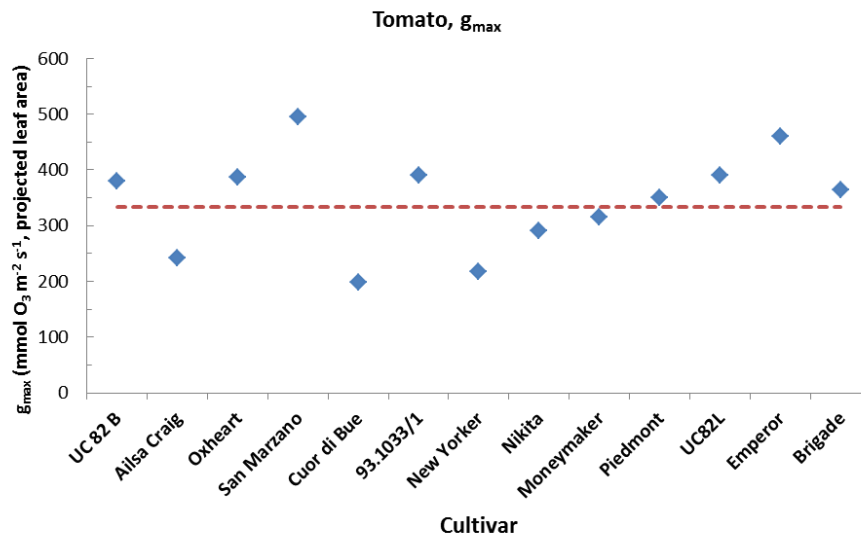


Figure 4.6 Derivation of g_{max} for tomato (see González-Fernández et al., 2014 for details).

f_{min}

The f_{min} value for tomato has been derived from the average of the values below the 5th percentile of all the g_{sto} measurements.

f_{phen}

The data used to establish the f_{phen} function for tomato are presented in **Figure 4.7**. g_{max} was assumed to occur at a fixed number of days since the start of the growing season.

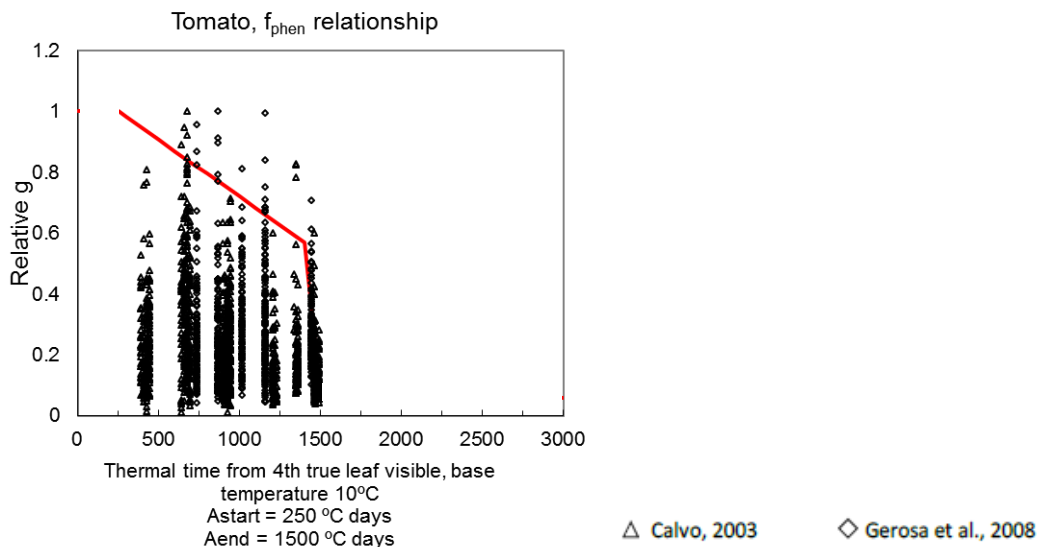


Figure 4.7 f_{phen} function for tomato (see González-Fernández et al., 2014 for details).

Table 4.3

Maximum stomatal conductance (g_{max}) values reported in field studies. Values were measured on sun exposed leaves under optimum environmental conditions for maximum stomatal opening (González-Fernández et al., 2014).

Reference	g _{max} (mmol O ₂ m ⁻² PLA s ⁻¹)	g _{max} derivation (mmol H ₂ O m ⁻² PLA s ⁻¹)	Country	Tomato cultivar	Time of day	Time of year	g _{max} measuring apparatus	Gas/leaf area	Growing conditions	Leaf
Bolaños and Hsiao, 1991	400	604 Values given in figure, 0.628, 0.642, 0.542 (mol m ⁻² s ⁻¹), 3 replicates	USA, California	UC82B	Midday	June to September	Binos Irga	H ₂ O/PLA	Field grown tomato leaves of high N status	Youngest expanded
Bolaños and Hsiao, 1991	359	542 Values given in figure, 0.542 (mol m ⁻² s ⁻¹), 3 replicates	USA, California	UC82B	–	June to September	Open system Irga	H ₂ O/PLA	Pots	Leaf age – 27 days old
Else et al., 1996	244	367 Values given in figure, 0.275, 0.246 (mol m ⁻² s ⁻¹), Each point a mean of 8 plants, 2 photoperiods	UK or The Netherlands	Ailsa Craig	Midday	–	Mk3, Dif porometer (Delta T)	H ₂ O/abaxial	Pots. Controlled environment rooms. Plants with 7 or 8 leaves	Fourth oldest leaf
Else et al., 1996	171	257 Values given in figure, 0.185, 0.185, 0.185, 0.175 (mol m ⁻² s ⁻¹), Each point a mean of 8 plants, 4 photoperiods	UK or The Netherlands	Ailsa Craig	Midday	–	Mk3, Diffusion porometer (Delta T)	H ₂ O/abaxial	Pots. Controlled environment rooms. Plants with 7 or 8 leaves	Fourth oldest leaf
Guidi et al., 2005	199	300 Values given in figure, 300 (mmol m ⁻² s ⁻¹), Mean of 3 replicates	Italy	Cuor di Bue	–	–	HCM-1000 Walz, IRGA	H ₂ O/PLA	Pots in Controlled environment chamber	2nd expanded leaf from the bottom
Guidi et al., 2005	390	588 Values given in figure, 587.5 (mmol m ⁻² s ⁻¹), Mean of 3 replicates	Italy	93.1033/1	–	–	HCM-1000 Walz, IRGA	H ₂ O/PLA	Pots in Controlled environment chamber	2nd expanded leaf from the bottom
Hao et al., 2000	217	327 Values given in table, 0.327 (mmol m ⁻² s ⁻¹) n = 36	Canada	New Yorker	10–14 h	–	Licor-6200, IRGA	H ₂ O/PLA	Pots in Controlled environment chamber	5th youngest leaf fully expanded
Patanè, 2011	364	549 Value given in figure, 0.95 cm s ₁	Italy	Brigade	13–14 h	Since May	Licor-1600, porometer	H ₂ O/abaxial	Field grown, well irrigated	Youngest expanded
Righi et al., 2012	460	693 Value given in text, 12 mm s ₁ , conversion factor given in text abaxial/adaxial ratio 2.88	Brazil	Emperor	11–14 h	March to May	Licor-1600, porometer	H ₂ O/abaxial	Bags in greenhouse	Top of canopy
This study (Gerosa et al., 2008)	387	584 Values given in Table. 425 (cm s ⁻¹), Percentile 95th from 700 measurements	Italy	Oxheart	8 to 17:30	July to September	AP4, Diffusion porometer (Delta T)	H ₂ O/PLA	Pots in OTC	Light exposed leaf
This study* (Gerosa et al., 2008)	496	748 Values given in Table. 825 (cm s ⁻¹), Percentile 95th from 700 measurements	Italy	San Marzano	8 to 17:30	July to September	AP4, Diffusion porometer (Delta T)	H ₂ O/PLA	Pots in OTC	Light exposed leaf
This study (Benifaíó**)	290	438 Calculated from measurements. Percentile 95th from 347 measurements	Spain	Nikita	8 to 17:00	August to September	HCM-1000 Walz	H ₂ O/PLA	Pots in OTC	Light exposed leaf
This study (Benifaíó)	313	472 Calculated from measurements. Percentile 95th from 253 measurements	Spain	Ailsa Craig	8 to 17:00	August to September	HCM-1000 Walz	H ₂ O/PLA	Pots in OTC	Light exposed leaf
This study (Benifaíó)	315	475 Calculated from measurements. Percentile 95th from 209 measurements	Spain	Money-maker	8 to 17:00	August to September	HCM-1000 Walz	H ₂ O/PLA	Pots in OTC	Light exposed leaf
This study (Benifaíó)	351	529 Calculated from measurements. Percentile 95th from 188 measurements	Spain	Piedmont	8 to 17:00	August to September	HCM-1000 Walz	H ₂ O/PLA	Pots in OTC	Light exposed leaf
This study (Benifaíó)	390	588 Calculated from measurements. Percentile 95th from 188 measurements	Spain	UC82L	8 to 17:00	August to September	HCM-1000 Walz	H ₂ O/PLA	Pots in OTC	Light exposed leaf
Mean = 334		SD = 92								
Range = 171–496		Median = 355								

* This study: González-Fernández et al., 2014; ** Benifaíó = site in Eastern Spain.

f_{light}

The data used to establish the f_{light} function for tomato are shown in **Figure 4.8**. The f_{light} modifying function was adjusted by boundary line analysis.

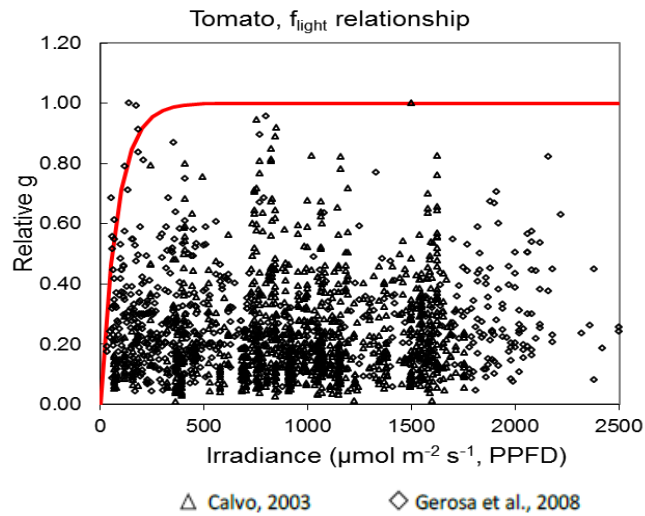


Figure 4.8 Derivation of f_{light} for tomato (see González-Fernández et al., 2014 for details).

f_{temp}

The data used to establish the f_{temp} function for tomato are shown in **Figure 4.9**. The f_{temp} modifying function was adjusted by boundary line analysis.

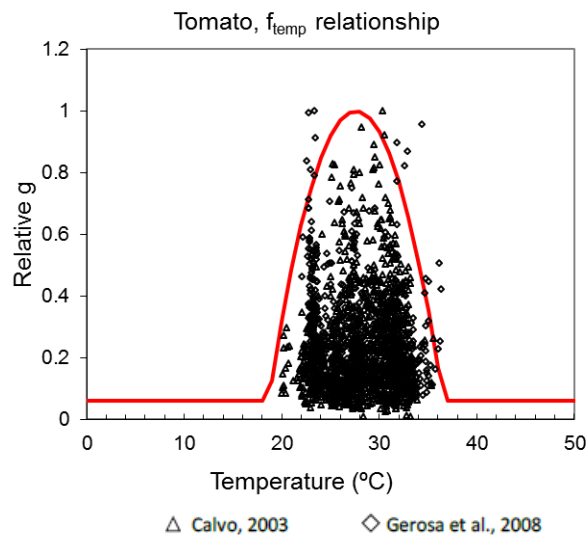


Figure 4.9 Derivation of f_{temp} for tomato (see González-Fernández et al., 2014 for details).

f_{VPD}

The data used to establish the f_{VPD} function for tomato are shown in **Figure 4.10**. The f_{VPD} modifying function was adjusted by boundary line analysis.

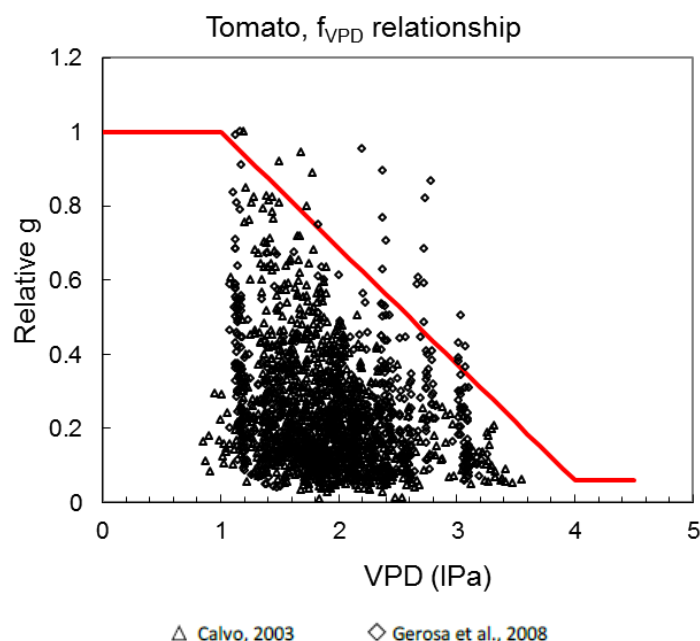


Figure 4.10 Derivation of f_{VPD} for tomato (see González-Fernández et al., 2014 for details).

f_{sw}

No limiting function for soil water content was considered for tomato since constant irrigation is provided during the whole growing period.

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5 Forest trees – Additional information on O₃ flux model parameterisation (Section III.3.5.3 of the manual)

5.1 Additional information on parameterisation of forest tree species and sources of information

The parameterisation of the O₃ stomatal flux model DO₃SE for risk assessment is generally based on data from mature tree species, except for Mediterranean deciduous oak species, for which the parameterisation is based on young trees of three oak species, i.e. *Quercus faginea* (Portuguese oak), *Q. pyrenaica* (Pyrenean oak) and *Q. robur* (pedunculate oak). Table III.11 in Chapter 3 of the Modelling and Mapping Manual describes a combined parameterisation for all three Mediterranean oak species. The parameterisations come from experiments used to derive critical levels and represent seedling physiology generally under well-watered conditions (Marzuoli et al., in prep.). Specific parameterisation of the individual species, and in the case of pedunculate oak application in different countries, is provided in **Table 5.1**. Parameterisation of the DO₃SE model using data from young oak tree species was considered to better represent risk assessment of deciduous trees in Mediterranean areas than the previous parameterisation based on mature Mediterranean beech.

Table 5.1 Parameterisation of the O₃ stomatal flux model for Mediterranean deciduous oak tree species (Marzuoli et al., in prep.).

Parameter	Units	Forest broadleaf deciduous tree species parameterisation - POD ₁ SPEC			
Region		Mediterranean			
Forest type		Broadleaf deciduous			
Species ⁱ	Common name	Pedunculate oak, Spain	Pedunculate oak, Italy	Pyrenean oak	Portuguese oak
	Latin name	<i>Quercus robur</i>	<i>Quercus robur</i>	<i>Quercus pyrenaica</i>	<i>Quercus faginea</i>
g_{max}	mmol O ₃ m ⁻² PLA s ⁻¹	235	235	310	280
f_{min}	fraction	0.13	0.13	0.13	0.13
light_a	-	0.006	0.006	0.006	0.006
T _{min}	°C	-5	-5	-5	-5
T _{opt}	°C	22	22	22	22
T _{max}	°C	35	35	35	35
VPD _{max}	kPa	1.1	1.1	1.1	1.1
VPD _{min}	kPa	3.1	3.1	3.1	3.1
ΣVPD _{crit}	kPa	-	-	-	-
PAW _t ⁱⁱ	%	-	-	-	-
SWC _{max} ⁱⁱ	% volume	- ⁽ⁱⁱⁱ⁾	- ⁽ⁱⁱⁱ⁾	- ⁽ⁱⁱⁱ⁾	- ⁽ⁱⁱⁱ⁾
SWC _{min} ⁱⁱ	% volume	- ⁽ⁱⁱⁱ⁾	- ⁽ⁱⁱⁱ⁾	- ⁽ⁱⁱⁱ⁾	- ⁽ⁱⁱⁱ⁾
SWP _{max}	MPa	- ⁽ⁱⁱⁱ⁾	- ⁽ⁱⁱⁱ⁾	- ⁽ⁱⁱⁱ⁾	- ⁽ⁱⁱⁱ⁾

SWP _{min}	MPa	- (iii)	- (iii)	- (iii)	- (iii)
f _{O3}	fraction	-	-	-	-
Astart_FD ^{iv}	day of year	Latitude model			
Aend_FD ^{iv}	day of year	Latitude model			
Leaf dimension	cm	4.0	5.0	5.5	2.5
Canopy height	m	25	25	15	20
f _{phen_a}	fraction	0.3	0.0	0.3	0.3
f _{phen_b}	fraction	(1.0)	(1.0)	(1.0)	(1.0)
f _{phen_c}	fraction	1.0	1.0	1.0	1.0
f _{phen_d}	fraction	(1.0)	(1.0)	(1.0)	(1.0)
f _{phen_e}	fraction	0.3	0.0	0.3	0.3
f _{phen_1_FD}	no. of days	50	20	50	50
f _{phen_2_FD}	no. of days	(200)	(200)	(200)	(200)
f _{phen_3_FD}	no. of days	(200)	(200)	(200)	(200)
f _{phen_4_FD}	no. of days	50	50	50	50
LIM _{start_FD}	year day	(0.0)	(0.0)	(0.0)	(0.0)
LIM _{send_FD}	year day	(0.0)	(0.0)	(0.0)	(0.0)

The values in brackets represent “dummy” values required for DO₃SE modelling purposes. “-” = parameterisation not required for this species.

ⁱ See Table 5.2 for source of parameterisations.

ⁱⁱ Soil water content (SWC) is calculated as the soil water content available for transpiration, i.e. the actual SWC minus the SWC at the wilting point. PAW_t is the threshold for plant available water (PAW), above which stomatal conductance is at a maximum.

ⁱⁱⁱ The parameterisation for these tree species were predominantly derived from experiments that were conducted under non-limiting water supply. Hence, non-limiting soil water condition should be assumed, represented in the flux model by f_{SWP} = 1.

^{iv} Latitude function, see section 5.3; f_{temp} = growing season is assumed to occur when air temperatures fall within T_{min} and T_{max} thresholds of the f_{temp} relationship. Use of actual data is recommended if available.

Table 5.2 provides reference to the literature where further details of some aspects of the parameterisation of the DO₃SE model for tree species can be found.

Table 5.2 *References for the parameterisation of the DO₃SE model for all tree species given in Table III.11 in Mapping Manual and Table 5.1 above.*

Parameter	References for forest tree species parameterisation					
Region	Boreal		Continental (Atlantic, Steppic, Pannonian)		Mediterranean	
Forest type	Coniferous	Broadleaf deciduous	Coniferous	Broadleaf deciduous	Broadleaf deciduous	Evergreen
Tree species	Norway spruce	Silver birch	Norway spruce	Beech	Deciduous oak species	Evergreen species
g_{max} [value – mmol O ₃ m ⁻² PLA s ⁻¹]	Hansson et al. (pers. comm.) [121]; Sellin (2001) [131]	Sellin & Kupper (2005) [239]; Uddling et al. (2005) [239]	Dixon et al. (1995) [131]; Emberson et al. (2000) [130]; Körner et al. (1979) [95]; Zweifel et al. (2000, 2001, 2002) [155]; Zimmermann et al. (1988) [126]	Keel et al. (2007) [184]; Körner et al. (1979) [152]; Kutsch et al. (2001) [331]; Matyssek et al. (2004) [132]; Nunn et al. (2005) [160]; Schaub (pers. comm.) [137]	Marzuoli et al. (in prep.)	Alonso et al. (2007) [199, 208]; Bussotti & Ferretti (2006) [174, 185, 209]; Castell et al. (1994) [192]; Corcuera et al. (2005) [146]; Damesin et al. (1998) [186]; Filho et al. (1998) [245]; Gratani et al. (2000) [172]; Infante et al. (1999) [351]; Manes et al. (1997) [398]; Mediavilla & Escudero (2003) [133]; Rhizopoulos & Mitrakos (1990) [272]; Sala & Tenhunen (1994) [180]; Tognetti et al. (1998) [212]
f_{min}	Hansson et al. (pers. comm.)	Uddling et al. (2005)	Körner (1994)	Körner (1994)	Marzuoli et al. (in prep.)	Alonso et al. (2007); Bussotti & Ferretti (2007)

light_a	Karlsson et al. (2000); Hansson et al. (pers. comm.)	Uddling et al. (2005)	Körner et al. (1995); Thoene et al. (1991); Zweifel et al. (2000, 2001, 2002)	Kutsch et al. (2001)	Marzuoli et al. (in prep.)	Alonso et al. (2007) ; Bussotti & Ferretti (2007)
T _{min} , T _{opt} , T _{max}	Hansson et al. (pers. comm.); Jarvis (1980); Karlsson et al. (2000); Lagergren & Lindroth (2002)	Uddling et al. (2005)	Braun et al. (pers. comm.); Zweifel et al. (2000, 2001, 2002)	Braun et al. (2010)	Marzuoli et al. (in prep.)	Bussotti & Ferretti (2007); Corcuera et al. (2005); Elvira et al. (2005); Manes et al. (1997); Ogaya & Peñuelas (2003); Tenhunen et al. (1987); Vitale et al. (2005)
VPD _{min} , VPD _{max}	Hansson et al. (pers. comm.); Zimmermann et al. (1988)	Uddling et al. (2005)	Braun et al. (pers. comm.); Zweifel et al. (2000, 2001, 2002)	Braun et al. (2010)	Marzuoli et al. (in prep.)	Alonso et al. (2007) ; Bussotti & Ferretti (2007); Elvira et al. (2005); Manes et al. (1997); Sala & Tenhunen (1994); Tognetti et al. (1998); Vitale et al. (2005)
SWP _{min} , SWP _{max}	Sellin (1997)	Uddling et al. (2005)	Zweifel et al. (2000, 2001, 2002)	Braun et al. (2010)	Marzuoli et al. (in prep.)	Acherar & Rambal (1992); Alonso et al. (2007); Castell et al. (1994); Elvira et al. (2005); Epron & Dreyer (1990); Pesoli et al. (2003); Rhizopoulos & Mitrakos (1990); Sala & Tenhunen (1992); Tognetti et al. (1998)

5.2 Estimating the start (A_{start_FD}) and end (A_{end_FD}) of the growing season using the latitude model

As described in Section III.3.5.3.1 of Chapter 3 of the Modelling and Mapping Manual, the EMEP latitude model (Simpson et al., 2012) is used to estimate the start (A_{start_FD}) and end (A_{end_FD}) of the growing season of many tree species. The model is mainly used due to its simplicity, but developments are ongoing to improve modelling the start and end of the growing season for tree species (see SBD-B). A_{start_FD} is defined as the date of leaf unfolding (deciduous & broadleaf evergreen species) or the start of leaf/needle physiological activity (coniferous and evergreen species). For species-specific flux models (POD_YSPEC), A_{start_FD} is estimated by the EMEP latitude model with the exception of temperate conifers south of ~55°N, where A_{start_FD} is defined by prevailing environmental conditions (using the f_{temp} function), and for Mediterranean trees where a year round growth period is assumed. A_{end_FD} is defined as the onset of dormancy; the EMEP latitude model is used to identify A_{end_FD} again with the exception of temperate conifers south of ~55°N where A_{end_FD} is defined by prevailing environmental conditions (using the f_{temp} function) and for Mediterranean trees where a year round growth period is assumed. The model gives a good agreement with observed phenological data for a range of deciduous species (birch, beech and oak; **Figure 5.1**); with measurements of carbon flux from CarboEurope (<http://www.carboeurope.org/>) which were used to identify the initiation and cessation of physiological activity; and was also able to describe the onset of forest green-up and dormancy as determined from European remotely sensed data (Zhang et al. 2004). However, there are uncertainties associated with the estimation of key growth periods (e.g. start of physiological activity and onset of dormancy in conifers, initiation of leaf flush and onset of leaf fall in deciduous species) for the “real” species.

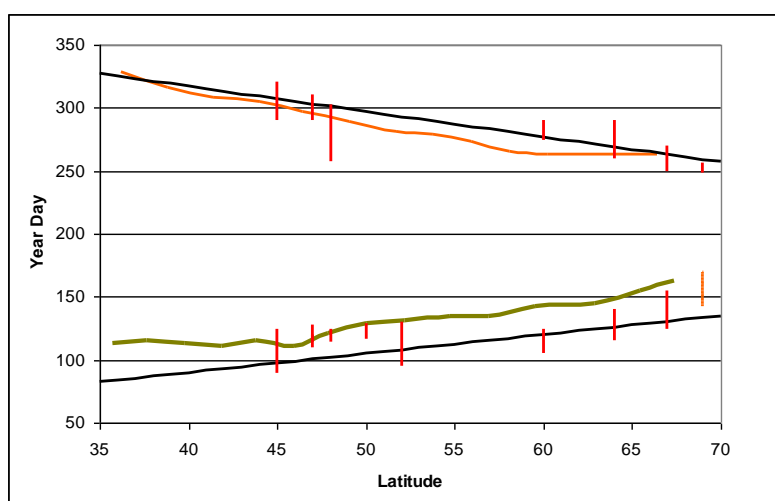


Figure 5.1 Comparison of observational phenological data with the EMEP latitude model. The black lines show the A_{start_FD} and A_{end_FD} determined from the EMEP latitude model. The green and orange lines show the onset of green-up and dormancy described by remotely sensed data for the year 2001 (Zhang et al. 2004) and the vertical red lines show the variation in observed dates for sites at specific latitudes for a number of different years.

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6 (Semi-)natural vegetation – Flux model parameterisation for selected (semi-)natural vegetation species and associated flux-effect relationships (Section III.3.5.4.2 of the manual)

6.1 Temperate perennial grasslands

Parameterisation of the O₃ stomatal flux model DO₃SE, sources of data and boundary lines were based on 5800 datapoints covering 26 semi-natural vegetation species from Boreal, Atlantic and Continental regions in Europe. Publications containing the experimental details of studies from where these were derived include Dawnay and Mills (2008), Hayes et al. (2006, 2010, 2011, 2012), Hewitt et al. (2016), Mills et al. (2009), Wagg et al. (2012, 2013), and Williamson et al. (2016). Additional data from unpublished studies were also included.

g_{max}

The g_{max} value was set at a fixed value of 100 mmol O₃ m⁻² PLA s⁻¹ for all species. This was to identify whether species with a high g_{max} were having a large influence on the dose-response relationships compared to those with low g_{max} as these would have proportionately higher accumulated flux.

f_{min}

The f_{min} value has been derived from measured values in experiments in solardomes at CEH-Bangor and is consistent with those derived for Mediterranean annual pasture.

f_{phen}

The f_{phen} function was set constant to one in recognition of the wide inter-annual variability in the duration and timing of the growing season of temperate perennial grassland species. It is recommended instead to use a three-month accumulation period within the maximum growth period, defined by the parameters A_{start} and A_{end} of Table III.13 in the manual.

f_{light}

The data used to establish the f_{light} function are shown in **Figure 6.1**. The f_{light} modifying function was set by boundary line analysis.

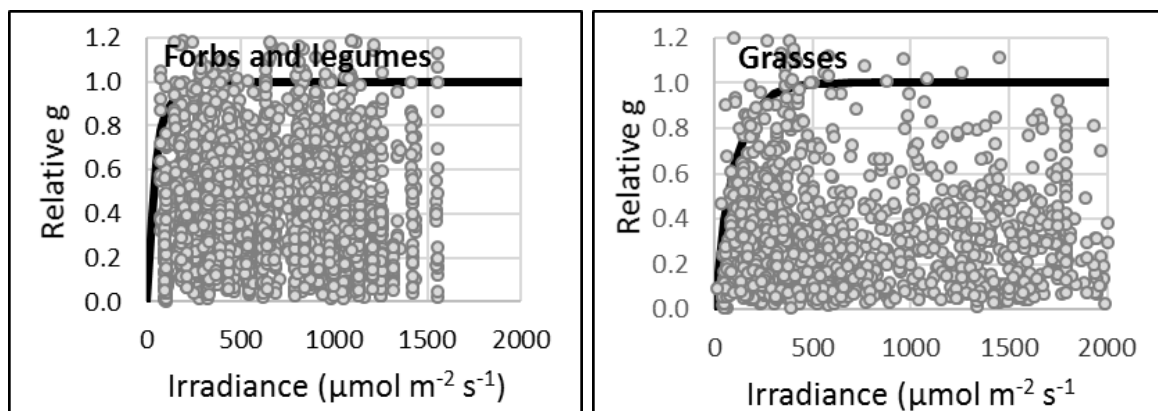


Figure 6.1 Derivation of f_{light} for forbs & legumes and grasses in temperate perennial grasslands.

The data used to establish the f_{temp} function are shown in **Figure 6.2**. The f_{temp} modifying function was set by boundary line analysis.

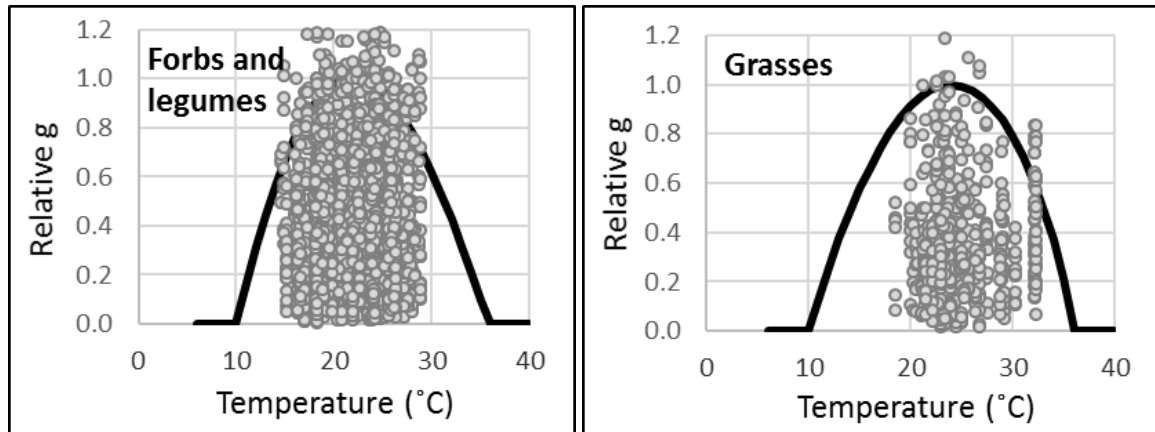


Figure 6.2 Derivation of f_{temp} for forbs & legumes and grasses in temperate perennial grasslands.

f_{VPD}

The data used to establish the f_{VPD} function are shown in **Figure 6.3**. The f_{VPD} modifying function was set by boundary line analysis.

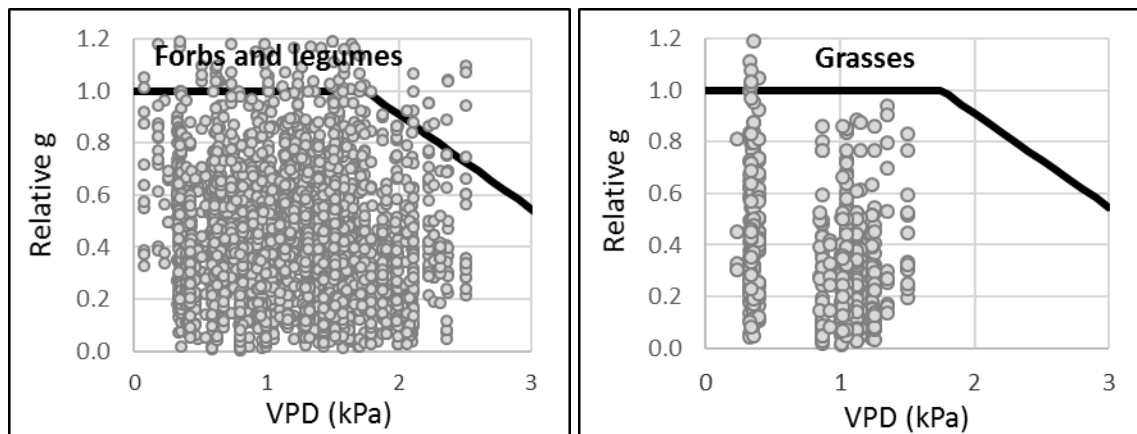


Figure 6.3 Derivation of f_{VPD} for forbs & legumes and grasses in temperate perennial grasslands.

f_{swc}

The data used to establish the f_{swp} function for are shown in **Figure 6.4**. The f_{swp} modifying function was set by boundary line analysis.

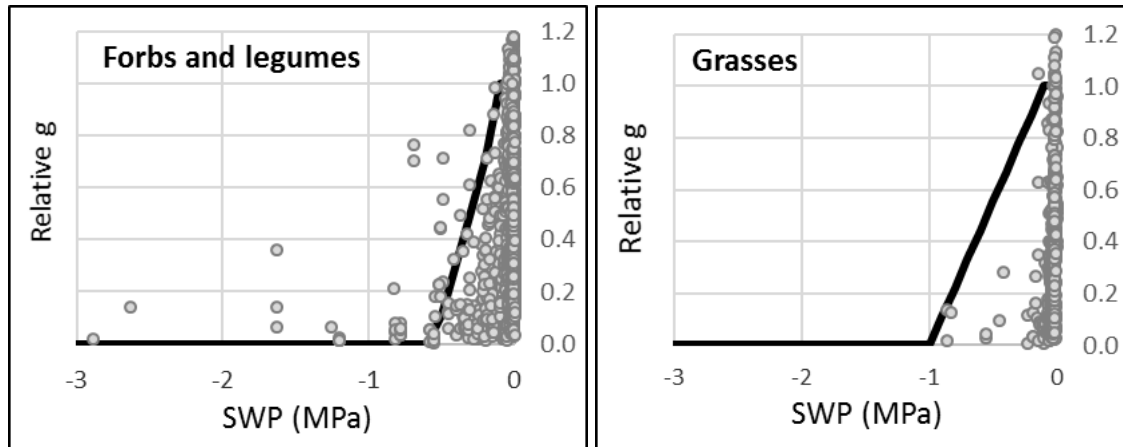


Figure 6.4 Derivation of f_{SWP} for forbs & legumes and grasses in temperate perennial grasslands.

6.2 Mediterranean annual pastures

Parameterisation of the O_3 stomatal flux model DO_3SE , sources of data and boundary lines were based on 159 datapoints measured in *Trifolium subterraneum* growing in natural pastures of central Spain. Publications containing the experimental details of studies from where these were derived include Alonso et al. (2007), González-Fernández et al. (2010), Sanz et al. (2016) and Calvete-Sogo et al. (2017).

The g_{max} value for legume species of Mediterranean annual pastures has been derived from field measurements on sun exposed leaves under optimum environmental conditions for stomatal opening (González-Fernández et al., 2010). g_{max} was calculated as the average of selected values above the 90th percentile from a dataset of stomatal conductance of *Trifolium subterraneum*, an ozone sensitive legume. This value is close to the average of g_{max} value (789 $mmol O_3 m^{-2} PLA s^{-1}$) derived for annual Mediterranean species in the same study and falls within the range of g_{max} values measured in other annual *Trifolium* species (749 - 895 $mmol O_3 m^{-2} PLA s^{-1}$) in an independent study under Mediterranean conditions (Calvete-Sogo et al., 2017).

f_{min}

The f_{min} value for Mediterranean annual pastures has been derived from measured values in the field as described in González-Fernández et al. (2010).

f_{phen}

The f_{phen} function for Mediterranean annual pastures was set constant to one (Sanz et al., 2016) in recognition of the wide inter-annual variability in the duration and timing of the growing season of Mediterranean annual pastures (Peco et al., 2009; González-Fernández et al., 2010). Instead, it is recommended to use a 45-day long accumulation period within the maximum growth period in spring (Sanz et al., 2016), defined by the parameters A_{start} and A_{end} of Table III.13 of the manual or by in situ observations for local applications.

f_{light}

The data used to establish the f_{light} function for Mediterranean annual pastures are shown in **Figure 6.5**. The f_{light} modifying function was adjusted by boundary line analysis.

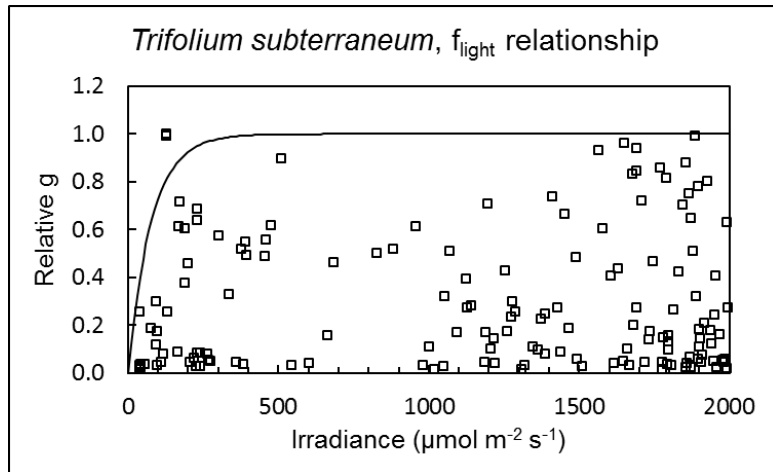


Figure 6.5 Derivation of f_{light} for Mediterranean annual pastures. Reprinted from *Atmospheric Environment* 44, Gonzalez-Fernandez, I., Bermejo, V., Elvira, S., Sanz, J., Gimeno, B.S., Alonso, R., *Modelling annual pasture dynamics: application to stomatal ozone deposition*, 2507-2517, Copyright 2010, with permission from Elsevier.

f_{temp}

The data used to establish the f_{temp} function for Mediterranean annual pasture are shown in **Figure 6.6**. The f_{temp} modifying function was adjusted by boundary line analysis.

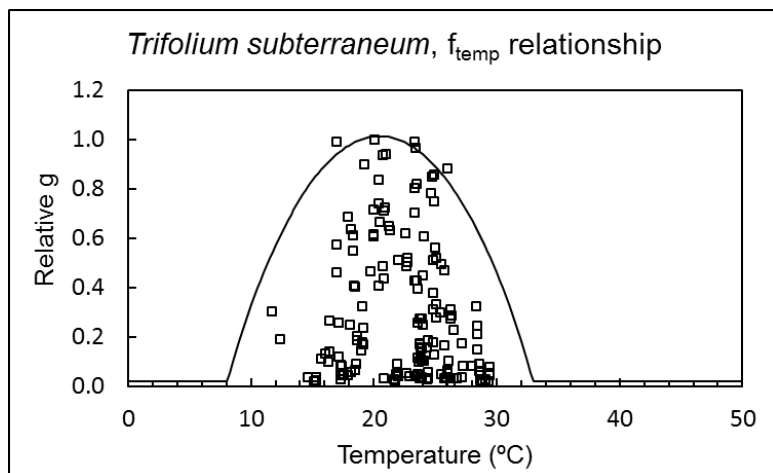


Figure 6.6 Derivation of f_{temp} for Mediterranean annual pasture. Reprinted from *Atmospheric Environment* 44, Gonzalez-Fernandez, I., Bermejo, V., Elvira, S., Sanz, J., Gimeno, B.S., Alonso, R., *Modelling annual pasture dynamics: application to stomatal ozone deposition*, 2507-2517, Copyright 2010, with permission from Elsevier.

f_{VPD}

The data used to establish the f_{VPD} function for Mediterranean annual pasture are shown in **Figure 6.7**. The f_{VPD} modifying function was adjusted by boundary line analysis.

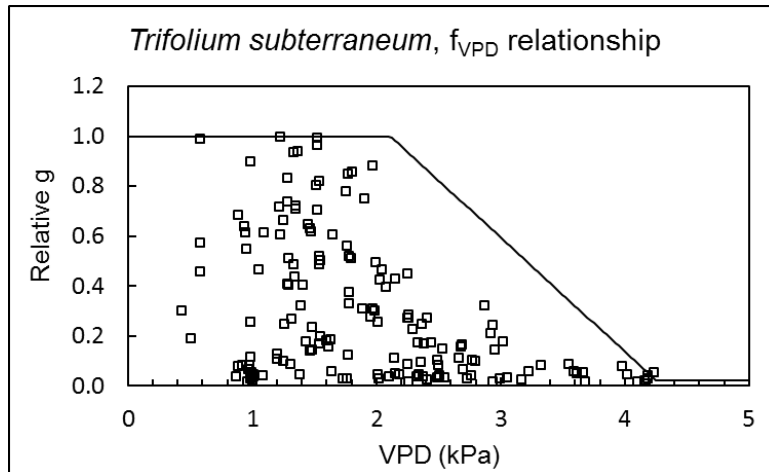


Figure 6.7 Derivation of f_{VPD} for Mediterranean annual pasture. Reprinted from *Atmospheric Environment* 44, Gonzalez-Fernandez, I., Bermejo, V., Elvira, S., Sanz, J., Gimeno, B.S., Alonso, R., *Modelling annual pasture dynamics: application to stomatal ozone deposition*, 2507-2517, Copyright 2010, with permission from Elsevier.

f_{swc}

The data used to establish the f_{swc} function (% volumetric water content) for Mediterranean annual pasture are shown in **Figure 6.8**. The f_{swc} modifying function was adjusted by boundary line analysis.

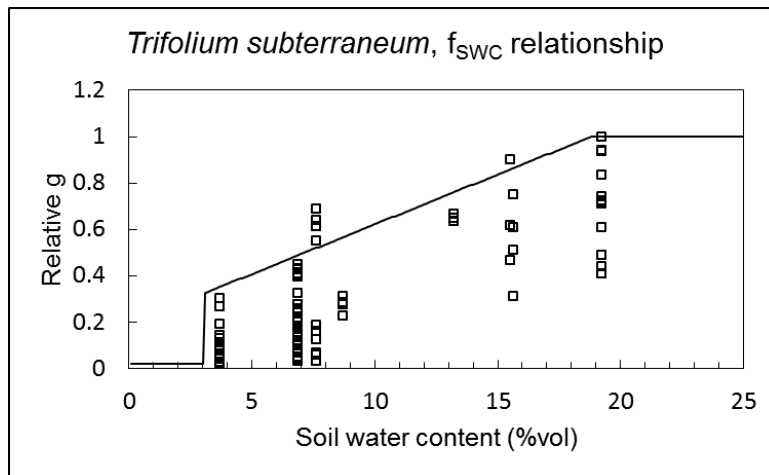


Figure 6.8 Derivation of f_{swc} for Mediterranean annual pasture. Reprinted from *Atmospheric Environment* 44, Gonzalez-Fernandez, I., Bermejo, V., Elvira, S., Sanz, J., Gimeno, B.S., Alonso, R., *Modelling annual pasture dynamics: application to stomatal ozone deposition*, 2507-2517, Copyright 2010, with permission from Elsevier.

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7 AOT40-based critical levels and response functions for O₃ (Section III.3.7 of the manual)

7.1 Introduction

These are based on accumulation of the hourly mean O₃ concentration at the top of the canopy over a threshold concentration of 40 ppb during daylight hours (when global radiation is more than 50 W m⁻²) for the appropriate time-window (AOT40) and thus do not take account of the stomatal influence on the amount of O₃ entering the plant. Hence, the spatial distribution of the risk of adverse impacts on vegetation generally mimics the spatial distribution of O₃ concentration and is different from the spatial distribution of POD_γ on the pan-European scale (Mills et al., 2011, Simpson et al., 2007). AOT40-based critical levels are suitable for estimating the risk of damage where climatic data or suitable flux models are not available and/or areas where no climatic or water restrictions to stomatal O₃ flux are expected. Economic losses should not be estimated using this method.

Note: The text presented here is reproduced from earlier versions of Chapter 3, with minor editorial changes. Since the O₃ critical levels workshop in Obergurgl in November 2005 only the AOT40-based critical level for horticultural crops has been revised at the 28th ICP Vegetation Task Force meeting in Rome in February 2015. The AOT40-based critical levels were discussed at the most recent critical levels workshop (Madrid, November 2016) and 30th ICP Vegetation Task Force Meeting (Poznan, February 2017) and no changes were recommended to the critical levels for growth, yield or biomass effects. The critical level for visible injury was removed from the manual as there was no longer sufficient scientific support for its inclusion.

7.2 Crops

AOT40-based critical levels and response functions

Agricultural crops

The concentration-based critical level for agricultural crops has been derived from a linear relationship between AOT40 and relative yield for wheat, developed from the results of open-top chamber experiments conducted in Europe and the USA (**Table 7.1; Figure 7.1**). Newer data (Gelang et al., 2000) has been added to that derived by Fuhrer et al. (1997) and quoted in the earlier version of the Mapping Manual (UNECE, 1996). Thus, the critical level for wheat is based on a comprehensive dataset including 9 cultivars. The AOT40 corresponding to a 5% reduction in yield is 3.3 ppm h (95% Confidence Interval range 2.3-4.4 ppm h). This value has been rounded down to 3 ppm h for the critical level. The critical level for agricultural crops is only applicable when nutrient supply and soil moisture are not limiting, the latter because of sufficient precipitation or irrigation (Fuhrer, 1995).

The time period over which the AOT40 is calculated should be three months and the timing should reflect the period of active growth of wheat and be centred around. For further explanation, please see Mills et al. (2007). It is not recommended that exceedance of the concentration-based critical level for agricultural crops is converted into economic loss; it should only be used as an indication of ecological risk (Fuhrer, 1995).

Horticultural crops

This text was prepared following decisions made at the 2015 Task Force meeting of the ICP Vegetation based on evidence published in González-Fernández et al. (2014). A concentration-based critical level has been derived for horticultural crops that are growing with adequate nutrient and water supply. An AOT40 of 8.4 ppm h (95% Confidence Interval range 1.2-15.6 ppm h) is equivalent to a 5% reduction in fruit yield for tomato, and has been derived from an exposure-response function developed from a comprehensive dataset including 5 O₃-sensitive cultivars ($r^2 = 0.63$, $p < 0.001$, **Figure 7.2**, see González-Fernández et al., 2014). This value has been rounded down to 8 ppm h for the critical level. Although

statistical analysis has indicated that water melon may be more sensitive to O₃ than tomato, the dataset for water melon is not sufficiently robust for use in the derivation of a critical level because the data is only for one cultivar (Mills et al., 2007). Tomato is considered suitable for the derivation of the critical level since it is classified as a moderately-sensitive crop and a suitably robust function is available. It is also an economically important horticultural crop. Other horticultural crops such as lettuce and bean are more sensitive but have a less robust response-function. The data used in the derivation of the critical level for horticultural crops is from experiments conducted in Spain and Italy. The time period for accumulation of AOT40 is three months, starting at the 4th true leaf stage (BBCH code = 14) which is normally the date of planting (see Section III.3.5.2.1 of the manual for guidance). It is not recommended that the exceedance of the concentration-based critical level for horticultural crops is converted into economic loss; it should only be used as an indication of ecological risk during the most sensitive environmental conditions (Fuhrer, 1995).

The AOT40-based critical levels for crops and the response functions from which they were derived are presented in **Table 7.1** and **Figures 7.1** and **7.2**.

Table 7.1 *AOT40-based critical levels and response functions for agricultural and horticultural crops.*

Category	Agricultural	Horticultural
Representative crop	Wheat	Tomato
Yield parameter	Grain yield	Fruit yield
% reduction for critical level	5 %	5 %
Critical level (AOT40, ppm h)	3	8
Countries involved in experiments	Belgium, Finland, Italy, Sweden	Spain, Italy
Number of data points	52	17
Number of cultivars	9	5 ozone-sensitive cultivars
Data sources	Mills et al., 2007	González-Fernández et al., 2014
Time period	3 months	3 months
Response function	$RY = 0.99 - 0.0161 * AOT40$	$RY = 1.01 - 0.0069 * AOT40$
r ²	0.89	0.63
P value	< 0.001	< 0.001

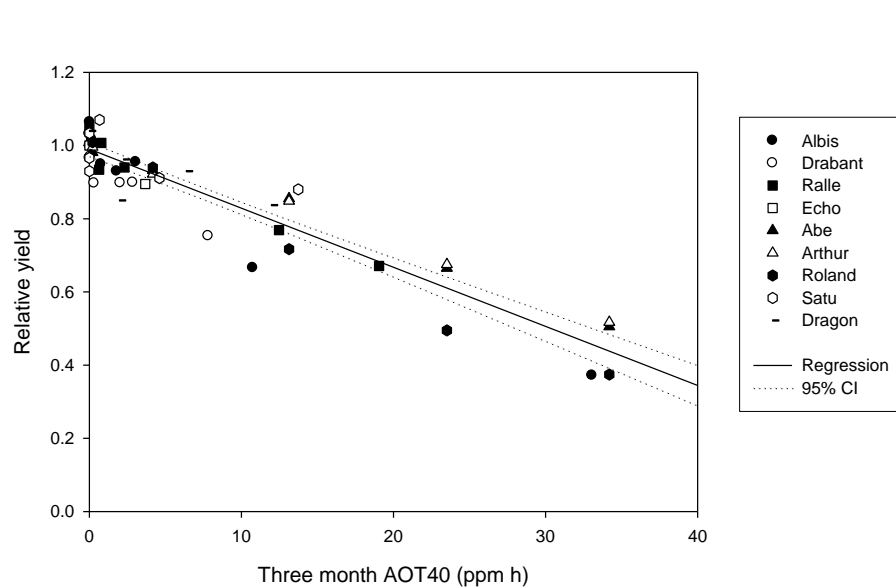


Figure 7.1 *Wheat yield-response function used to derive the concentration-based critical levels for agricultural crops ($r^2 = 0.89$) (data from Fuhrer et al., 1997 and Gelang et al., 2000, reproduced in Mills et al., 2007). Dotted lines represent 95% confidence intervals.*

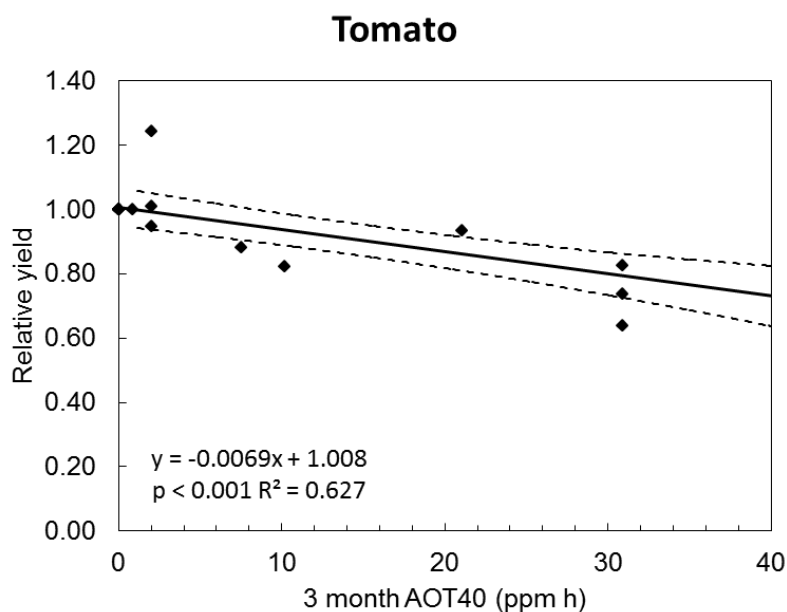


Figure 7.2 *Tomato yield-response function for O_3 -sensitive cultivars used to derive the concentration-based critical levels for horticultural crops ($r^2 = 0.63$, $p < 0.001$) (González-Fernández et al., 2014). Dotted lines represent 95% confidence intervals.*

Calculating exceedance of the AOT40-based critical levels

Step 1: Determine the accumulation period

Agricultural crops

The timing of the three month accumulation period for agricultural crops should reflect the period of active growth of wheat and be centred on the timing of anthesis. A survey of the development of winter wheat conducted at 13 sites in Europe by ICP Vegetation participants in 1997 and 1998, revealed that anthesis can occur as early as 2 May in Spain and as late

as 3 July in Finland (Mills and Ball, 1998, Mills et al., 2007). Thus, a risk assessment for O₃ impacts on crops would benefit from the use of a moving time interval to reflect the later growing seasons in northern Europe. For guidance, default time periods have been provided for five geographical regions as indicated in **Table 7.2**.

Horticultural crops

The timing of the start of the growing season is more difficult to define because horticultural crops are repeatedly sown over several months in many regions especially in the Mediterranean area. For local application within Mediterranean countries, appropriate 3 month periods should be selected between March and August for eastern Mediterranean areas, and March and October the Western Mediterranean areas. Since the cultivars used to derive the response function for tomato also grow in other parts of Europe, it is suggested that appropriate 3 month periods are selected between the period April to September for elsewhere in Europe

Steps 2 and 3: Determine the O₃ concentration at the top of the canopy.

The O₃ concentration at the canopy height can be calculated using the methods described in Section III.3.4.2 of the Modelling and Mapping Manual. For agricultural crops the default height of the canopy is 1 m whilst for horticultural crops (represented by tomato) it is 2m.

Step 4: Continue as described in Section III.3.7.2 of the Modelling and Mapping Manual.

Table 7.2 *Regional classification of countries for default time periods for calculation of AOT40 for agricultural crops. See text for time periods for horticultural crops.*

Region	Abbreviation	Three month time period	Possible default countries
Eastern Mediterranean	EM	1 March to 31 May	Albania, Bosnia and Herzegovina, Bulgaria, Croatia, Cyprus, Greece, FYR Macedonia, Malta, Montenegro, Serbia, Slovenia, Turkey
Western Mediterranean	WM	1 April to 30 June	Holy See, Italy, Monaco, Portugal, San Marino, Spain
Continental Central Europe	CCE	15 April to 15 July	Armenia, Austria, Azerbaijan, Belarus, Czech Republic, France ¹ , Georgia, Germany, Hungary, Kazakhstan, Kyrgyzstan, Liechtenstein, Republic of Moldova, Poland, Romania, Russian Federation, Slovakia, Switzerland, Ukraine
Atlantic Central Europe	ACE	1 May to 31 July	Belgium, Ireland, Luxembourg, Netherlands, United Kingdom
Northern Europe	NE	1 June to 31 August	Denmark, Estonia, Finland, Iceland, Latvia, Lithuania, Norway, Sweden

¹ As an average between Western Mediterranean and Atlantic Central Europe

7.3 Forest trees

Scientific basis

The experimental database for trees that was first presented at the UNECE Workshop in Gothenburg 2002 was re-analysed for the Obergurgl Workshop (2005) and expanded to include additional correlations with AOT20, AOT30, and AOT50 (Karlsson et al., 2003). Furthermore, the tree species included in the analysis have been separated into four species categories (**Table 7.3**) based on the sensitivity of growth responses to O₃. It should be emphasised that this categorisation is based on growth as a measure of effect. As a result of this differentiation of species, linear regressions between exposure and response have the highest r² values (**Table 7.4**).

Table 7.3 Sensitivity classes for the tree species based on effects of O₃ on growth (Karlsson et al., 2003).

Ozone-sensitive species		Moderately ozone-sensitive species	
Deciduous	Coniferous	Deciduous	Coniferous
<i>Fagus sylvatica</i> <i>Betula pendula</i>	<i>Picea abies</i> <i>Pinus sylvestris</i>	<i>Quercus petraea</i> , <i>Quercus robur</i>	<i>Pinus halepensis</i>

Using the sensitivity categories described above, AOT40 gave the highest r² values of the AOTX indices tested (**Figure 7.3**). However, the difference between the r² values for AOT40 and AOT30 was small (0.62 and 0.61 respectively for the combined birch and beech dataset, Table 7.4).

Based on the analysis described in Table 7.4, the concentration-based critical level of O₃ for forest trees was reduced from an AOT40 value of 10 ppm h (Kärenlampi & Skärby, 1996) to 5 ppm h (range 1-9 ppm h, determined by the 99% confidence intervals), accumulated over one growing season (Figure 7.3). This value of 5 ppm h is associated with a 5% growth reduction per growing season for the deciduous sensitive tree species category (beech and birch). The 5 % growth reduction was clearly significant as judged by the 99% confidence intervals. This increase in the robustness of the dataset and the critical level represents a substantial improvement compared to the 10% growth reduction associated with the previous O₃ critical level of an AOT40 of 10 ppm h (Kärenlampi & Skärby, 1996). Furthermore, it represents a continued use of sensitive, deciduous tree species to represent the most sensitive species under most sensitive conditions. As previously, it should be strongly emphasized that these values should not be used to quantify O₃ impacts for forest trees under field conditions. Further information can be found in Karlsson et al. (2004).

Observation of visible injury in young trees in ambient air at Lattecaldo, in southern Switzerland has shown that a reduction of the O₃ critical level to 5 ppm h AOT40 would also protect the most sensitive species from visible injury (Van der Hayden et al., 2001, Novak et al., 2003). Furthermore, Baumgarten et al. (2000) detected visible injury on the leaves of mature beech trees in Bavaria well below 10 ppm h AOT40.

Table 7.4 Statistical data for regression analysis of the relationship between AOTX O₃ exposure indices (in ppm h) and percentage reduction of total and above-ground biomass for different tree species categories (Karlsson et al., 2003).

Ozone index/plant category	Linear regression			
	r ²	p for the slope	p for the intercept	slope
<i>AOT20</i>				
Birch, beech	0.52	<0.01	0.70	- 0.357
Oak	0.57	<0.01	0.73	- 0.142
Norway spruce, Scots pine	0.73	<0.01	0.31	- 0.086
<i>AOT30</i>				
Birch, beech	0.61	<0.01	0.63	- 0.494
Oak	0.61	<0.01	0.79	- 0.170
Norway spruce, Scots pine	0.76	<0.01	0.61	- 0.110
<i>AOT40</i>				
Birch, beech	0.62	<0.01	0.31	- 0.732
Oak	0.65	<0.01	0.73	- 0.216
Norway spruce, Scots pine	0.79	<0.01	0.86	- 0.154
<i>AOT50</i>				
Birch, beech	0.53	<0.01	0.05	- 1.033
Oak	0.62	<0.01	0.82	- 0.248
Norway spruce, Scots pine	0.76	<0.01	0.16	- 0.188

AOT40-based critical levels and response functions

The AOT40-based critical level for forest trees is presented in **Table 7.5**, with the associated function presented in **Figure 7.3**.

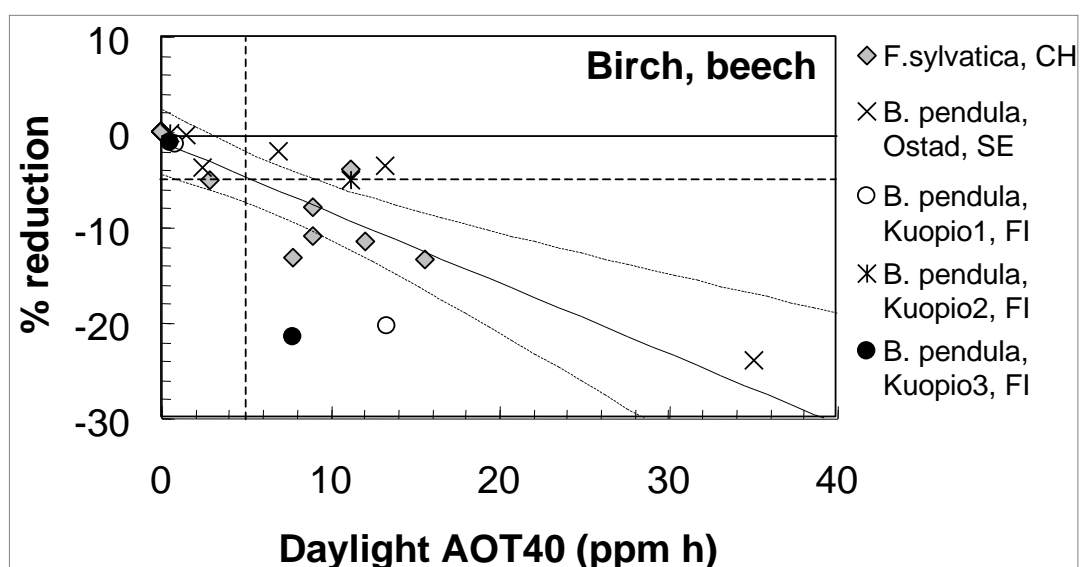


Figure 7.3 The relationship between percentage reduction in biomass and AOT40, on an annual basis, for the deciduous, sensitive tree species category, represented by beech and birch. The relationship was analysed by linear regression with 99% confidence intervals. Explanations for the figure legends can be found in Karlsson et al. (2003).

Table 7.5 AOT40-based critical levels and response functions for forest trees.

Category	Deciduous trees
Representative species	Birch and beech species
Effect parameter	Whole tree biomass
% reduction for critical level	5%
Critical level (AOT40, ppm h)	5 ppm h
Countries involved in experiments	Sweden, Finland, Switzerland
Number of data points	21
Data sources	Karlsson et al. (2007)
Time period	Growing season
Response function	Based on annual reduction in total tree biomass
r^2	0.62
P value	<0.01

Calculating exceedance of the AOT40-based critical level

The method described in Section III.3.7.2 of the Modelling and Mapping Manual should be followed incorporating the following recommendations specific for forest trees:

Step 1: The default exposure window for the accumulation of AOT40 is suggested to be 1 April to 30 September for all deciduous and evergreen species in all regions throughout Europe. This time period does not take altitudinal variation into account and should be viewed as indicative only. It should be stressed that it should only be used where local information is not available. When developing local exposure windows, the following definitions should be used:

- Onset of growing season in deciduous species: the time at which flushing has initiated throughout the entire depth of crown.
- Cessation of growing season in deciduous species: the time at which the first indication of autumn colour change is apparent.
- Onset of growing season in evergreen species: when the night temperatures are above -4°C for 5 days: if they do not fall below -4°C , the exposure window is continuous.
- Cessation of growing season in evergreen species: when the night temperatures are below -4°C for 5 days: if they do not fall below -4°C , the exposure window is continuous.

Steps 2 and 3: It is important that the calculation of AOT40 is based on O_3 concentrations at the top of the canopy as described in Section III.4.2. The suggested default canopy height for forest trees is 20m.

Step 4: Continue as described in Section III.3.7.2 in the Modelling and Mapping Manual.

7.4 (Semi-)natural vegetation

Scientific background and AOT40-based critical levels

The critical levels for (semi-)natural vegetation (**Table 7.6**) are applicable to all sensitive semi-natural vegetation and natural vegetation, excluding forest trees and woodlands, described here collectively as (semi-)natural vegetation. Two AOT40-based critical levels were agreed at the Obergurgl (2005) workshop.

Table 7.6 Summary of AOT40-based critical levels for (semi-)natural vegetation.

(Semi-) natural vegetation dominated by:	Critical level	Time period	Effect
Annuals	An AOT40 of 3 ppm h	3 months (or growing season, if shorter)	Growth reduction and/or seed production reduction in annual species
Perennials	An AOT40 of 5 ppm h	6 months	Effects on total above-ground or below-ground biomass and/or on the cover of individual species and/or on accelerated senescence of dominant species

Critical level for effects on communities of (semi-)natural vegetation dominated by annuals

The criteria for adverse effects on (semi-)natural vegetation communities dominated by annuals are effects on growth and seed production for annual species. This critical level is based on statistically significant effects or growth reductions of greater than 10% on sensitive taxa of grassland and field margin communities. The value of 3 ppm h is sufficient to protect the most sensitive annuals. In contrast to crops and tree species, only limited experimental data are available for a small proportion of the vast range of species found across Europe. This means that analysis of exposure-response data for individual species to derive a critical level value is more difficult. Instead, the recommended critical level is based on data from a limited number of sensitive species. The value of 3 ppm h was originally proposed at the Kuopio workshop (Kärenlampi & Skärby, 1996) and confirmed at the Gerzensee workshop (Fuhrer & Achermann, 1999) and subsequently at the Obergurgl workshop (Wieser and Tausz, 2006). At the time of the Kuopio workshop, no exposure-response studies were available for derivation of a critical level for (semi-)natural vegetation, based on a 10% response. Instead, data from field-based experiments with control and O₃ treatments were used to identify studies showing significant effects at relatively low O₃ exposures. **Table 7.7** summarises the key field chamber and field fumigation experiments which supported the original proposal of this critical level for (semi-)natural vegetation.

A number of studies have clearly demonstrated that the effects of O₃ in species mixtures may be greater than those on species grown alone or only subject to intraspecific competition. Therefore, the critical level needs to take into account the possibility of effects of interspecific competition in reducing the threshold for significant effects; indeed three of the four experiments listed in Table 7.6 include such competitive effects. By the time of the Gothenburg workshop (2002), the most comprehensive study of O₃ effects on species mixtures involving species which are representative of different communities across Europe, is the EU-FP5 BIOSTRESS (BIODiversity in Herbaceous Semi-Natural Ecosystems Under STRESS by Global Change Components) programme. Results to date from the BIOSTRESS programme, including experiments with species from the Mediterranean dehesa community, indicate that exposures to O₃ exceeding an AOT40 of around 3 ppm h may cause significant negative effects on annual and perennial plant species (see Fuhrer et al., 2003). The BIOSTRESS mesocosm experiments with two-species mixtures indicated

that exposures during only 4-6 weeks early in the growing season may cause shifts in species balance. The effect of this early stress may last for the rest of the growing season.

Table 7.7 Summary of key experiments supporting the critical level of 3 ppm h (now adopted for communities dominated by annuals), as proposed at the Kuopio workshop (Ashmore & Davison, 1996).

Species or community	Most sensitive species	AOT40 (ppm h)	Response	Reference
Individual plants	<i>Solanum nigrum</i>	4.2	-23%; shoot mass	Bergmann et al., 1996
	<i>Malva sylvestris</i>	3.9	-54%; seed mass	
Mesocosms of four species	<i>Trifolium repens</i>	5.0	-13%; shoot mass	Ashmore & Ainsworth, 1995
Mesocosms of seven species	<i>Festuca ovina</i>	7.0	-32%; shoot mass	Ashmore et al., 1996
	<i>Leontodon hispidus</i>	7.0	-22%; shoot mass	
Ryegrass-clover sward	<i>Trifolium repens</i>	5.0	-20%; shoot mass	Nussbaum et al., 1995

The key experiments from the BIOSTRESS programme, which support the proposed critical level for communities dominated by annual species, are summarised in **Table 7.8**. Taken together, these studies support a critical level in the range 2.5-4.5 ppm h, with a mean value of 3.3 ppm h.

The value of the critical level for communities dominated by annuals is further supported by other published data, for instance, for wetland species by Power and Ashmore (2002) and Franzaring et al. (2000) and for dehesa pastures by Sanz et al. (2016). In individual plants from wild strawberry populations growing at high latitudes, Manninen et al. (2003) observed a significant biomass decline of >10% at 5 ppm h from June-August.

Table 7.8 Summary of experiments from the BIOSTRESS programme which support the recommended critical level for communities dominated by annuals (reviewed by Fuhrer et al., 2003).

Responsive species	Competitor species	Variable showing significant response	Corresponding AOT40	Reference
<i>Trifolium pratense</i>	<i>Poa pratensis</i>	Biomass (-10%)	4.4 ppm h*	Gillespie & Barnes, unpublished data
<i>Veronica chamaedrys</i>	<i>Poa pratensis</i>	Species biomass ratio	3.6 ppm h	Bender et al. (2002)
<i>Trifolium cherleri</i> , <i>T. striatum</i>	<i>Briza maxima</i>	Flower production	2.2-2.7 ppm h	Gimeno et al. (2003); Gimeno et al. (2004).
<i>Trifolium cherleri</i>	<i>Briza maxima</i>	Seed output	2.4 ppm h	Gimeno et al. (2004)

* Estimated from exposure-response functions

Critical level for effects on communities dominated by perennial species

A critical level of an AOT40 of 5 ppm h over 6 months to prevent adverse effects in communities dominated by perennial species was recommended at the Obergurgl workshop (2005). Since this critical level is based on average AOT40 values in experiments with a duration of several years, mapping of exceedance of this critical level should be based on 5-year mean values of AOT40.

Table 7.9 Key findings of the studies used to establish the new critical level.

Location and community	Duration and AOT40 in control (C) and lowest effect treatment (T)	Biomass data	Species response
Southern Finland ¹ Dry grassland (7 species) in open-top chambers	3 years 1.7 ppm h (C) 8.5 ppm h (T) 3 months fumigation in summer of each year	40% and 34% reduction in above- and below- ground biomass, respectively; reduced N availability	64% reduction in biomass of <i>Campanula rotundifolia</i> ; 61% reduction in biomass of <i>Vicia cracca</i>
Wales ² Upland grassland (7 species) in solardomes	2 years 0 ppm h (C) 10, 12, 30 ppm h (T) 3 months fumigation each summer presented as current or 2050 background with/without episodic peaks	Significant increase in community senescence detected in 10 ppm h treatment; 7% reduction in cumulative above-ground community biomass in 30 ppm h treatment	Significant increase in senescence for <i>Festuca ovina</i> and <i>Potentilla erecta</i> at 10 ppm h; 15% reduction in <i>Anthoxanthum odoratum</i> biomass within the community at 30 ppm h
Northern England ³ Upland grassland Species-rich (11 species), in open-top chambers	18 months 3 ppm h (C). 10 ppm h (T) 6 months exposure during 'summer' to 50 ppb versus 30 ppb reducing to 35 ppb versus 20 ppb over 'winter'	16% reduction in total above-ground biomass	Significant reduction in biomass of <i>Briza media</i> and <i>Phleum bertolonii</i>
Southern England ⁴ Calcareous grassland (38 species) in open-top chambers	3 years 2.6 ppm h (C) 10.5, 13.3, 18.2 ppm h (T) Exposure for periods of 3-5 months each year at three levels of O ₃ , effects observed at lowest exposure	No significant effect on above-ground biomass	Significant change in community composition; loss of <i>Campanula rotundifolia</i>

Sources of data: ¹ Rämö et al. (2006); ² Mills et al. (2006); ³ Barnes & Samuelsson, quoted in Bassin et al. (2007); ⁴ Thwaites et al. (2006).

Note: (C) indicates AOT40 exposure in control treatment; (T) indicates AOT40 exposure in O₃ treatments. The AOT40 values above were calculated over a six-month period, even though the experimental period was in some cases shorter. For studies in which the fumigation period was less than six months, exposure outside the experimental period was added to both the control and treatment AOT40.

This new critical level is based on five studies that provide important new experimental evidence for the effects of O₃ on plant communities dominated by perennial species, in chamber and field fumigation studies. Four studies involved mesocosms established from seed, from plants taken from the field, or by transplanting communities from the field, while one study involved exposure to O₃ *in situ*. Because of the longer growth period of these communities, the AOT40 should be calculated over a six-month growth period. The response variables of perennial dominated communities include significant effects on total above-ground or below-ground biomass, on the cover of individual species and on accelerated senescence of dominant species. **Table 7.9** summarises the key findings of the studies which were used to establish the new critical level.

Calculating exceedance of the AOT40-based critical level

Follow the procedure outlined in Section III.3.7.2 of the Modelling and Mapping Manual using the following recommendations specific to (semi-)natural vegetation:

Step 1: Determine the accumulation period

Ideally, a variable time-window should be used in the mapping procedure to account for different growth periods of annuals and perennials in different regions of Europe. The AOT40 is calculated over the first three or six months of the growing season. The start of the growing season can be identified using:

1. Appropriate phenological models;
2. Information from local or national experts;
3. The default table below (**Table 7.10**).

For some species, the most relevant period of the growing season for O₃ effects may be less than three months in duration. In such cases, values of AOT40 should be calculated over the growing season, identified using appropriate local information.

Table 7.10 Default timing for the start and end of O₃ exposure windows for (semi-)natural vegetation. (Note: regional classifications of countries are suggested in Table 7.2.)

Region	Start date	End date	End date
		(annual-dominated communities)	(perennial-dominated communities)
Eastern Mediterranean*	1 March	31 May	31 Aug
Western Mediterranean*	1 March	31 May	31 Aug
Continental Central Europe	1 April	30 June	30 Sept
Atlantic Central Europe	1 April	30 June	30 Sept
Northern Europe	mid-April	mid-July	mid-October

* For mountain areas where the altitude is above 1500 m, use a start date of 1 April, with end dates of 30 June for annual-dominated communities and 30 September for perennial-dominated communities.

Steps 2 and 3: Determine the O₃ concentration at the top of the canopy

The AOT40 value should be calculated as the concentration at canopy height, using the information provided in Section III.3.4.2 of the Modelling and Mapping Manual and chapter 3 of this document. The transfer functions to make this calculation, based on deposition models, depend on a number of factors which may vary systematically between vegetation classes. These include canopy height and leaf area index (both natural variation and effects of management) and environmental variables such as vapour pressure deficit and soil moisture deficit. If such information is not available, it is recommended that the conversion factors for short grasslands are used as a default.

Steps 3 – 4: continue as indicated in III.3.7.2.

7.5 References

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